

1 Heavy metal accumulation and tolerance of energy grass (*Elymus elongatus*
2 subsp. *ponticus* cv. Szarvasi-1) grown in hydroponic culture

3 Gyula Sipos^a, Ádám Solti^b, Viktória Czech^b, Ildikó Vashegyi^c, Brigitta Tóth^d, Edit Cseh^b,
4 Ferenc Fodor^{b*}

5
6 ^a*Szent István University - Agricultural Research and Development Institute, Bikazug, Szarvas, H-5540, Hungary*

7 ^b*Department of Plant Physiology and Molecular Plant Biology, Eötvös University, Budapest, Pázmány P. lane*
8 *1/C, Budapest, H-1117, Hungary*

9 ^c *Department of Plant Molecular Biology, Agricultural Institute, Centre for Agricultural Research, HAS,*
10 *Brunszvik street 2. Martonvásár, H-2462, Hungary*

11 ^d*Department of Agricultural Botany and Crop Physiology, Institute of Crop Sciences, Centre for Agricultural*
12 *and Applied Economic Sciences, University of Debrecen, Böszörményi street 138. Debrecen, H-4032 Hungary,*

13

14

15

16 * *corresponding author*

17 Ferenc Fodor

18 e-mail: ffodor@ludens.elte.hu

19 phone/fax: +36 1381 2164

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21

22 **Abstract**

23

24 Phytoremediation is a plant based, cost effective technology to detoxify or stabilize
25 contaminated soils. Fast growing, high biomass, perennial plants may be used not only in
26 phytoremediation but also in energy production. Szarvasi-1 energy grass (*Elymus elongatus*
27 subsp. *ponticus* cv. Szarvasi-1), a good candidate for this combined application, was grown in
28 nutrient solution in order to assess its Cd, Cu, Ni, Pb and Zn accumulation and tolerance. Its
29 shoot metal accumulation showed the order Pb<Ni<Cu~Cd<Zn. In parallel with this, Pb and
30 Ni had no or very little influence on the growth, dry matter content, chlorophyll concentration
31 and transpiration of the plants. Cu and Cd treatment resulted in significant decreases in all
32 these parameters that can be attributed to Fe plaque formation in the roots suggested by
33 markedly increased Fe and Cu accumulation. This came together with decreased shoot and
34 root Mn concentrations in both treatments while shoot Cu and Zn concentrations decreased
35 under Cd and Cu exposure, respectively. Zn treatment had no effect or even slightly
36 stimulated the plants. This may be due to a slight stimulation of Fe translocation and a very
37 efficient detoxification mechanism. Based on the average 300 mg kg⁻¹ (dry mass) Zn
38 concentration which is 0.03% of the shoot dry mass the variety is suggested to be classified as
39 Zn accumulator.

40

41 Key words: heavy metal accumulation; iron plaque; phytoremediation; Szarvasi-1 energy
42 grass; tall wheatgrass; zinc accumulator

43

44 **1. Introduction**

45

46 Heavy metal contamination in soils is a worldwide environmental problem. The
47 contamination may be originated from natural and anthropogenic sources, the latter being
48 much more significant. Anthropogenic contamination may occur due to mining, industrial
49 activities, traffic, inadequate use of (phosphate) fertilisers in agriculture and amendment with
50 sewage sludge [1]. In Hungary, a recent environmental disaster underlines the significance of
51 the problem: an industrial accident at a caustic waste reservoir chain of the Ajkai Timföldgyár
52 alumina plant in Ajka, western Hungary in October 2010 flooded about 40 square kilometres
53 and two localities with alkaline wastes of metal containing sludge called red mud.

54 Heavy metals, naturally present or deposited in various concentrations, have different
55 solubility and mobility in the soil but may be mobilised and accumulated by plants [2]. This
56 means a major threat for heavy metal uptake by crop plants but also provide a possibility to
57 remove the metals from the soils by specific plant species. Phytoremediation techniques based
58 on naturally metal accumulating plants (accumulator plants) or chelate-assisted metal
59 mobilization and uptake, i.e. phytoextraction [3,4], may raise another problem of the fate of
60 harvested plant material. Fast growing, high biomass, perennial plants developed or
61 genetically designed for energy production may provide a feasible and cost-effective solution
62 [5].

63 Successful clean up of metals from the soil is based on the efficiency of plants to take
64 up and accumulate them in their roots and translocate them to the shoots whereas tolerance to
65 toxicity influences plant growth: large biomass also provides higher capacity for storage.
66 These processes depend on various transport proteins the presence and function of which
67 should be taken into account when new species are characterised.

68 Szarvasi-1 energy grass (*Elymus elongatus* subsp. *ponticus* cv. Szarvasi-1) was bred
69 from a native population of tall wheatgrass in Hungary that was adapted to slightly salty
70 habitats [6,7]. It has a fibrous root system that may reach 3.5 m whereas the shoot may grow
71 to 1.8-2.2 m. In spite of its high biomass yield it is well adapted to drought, flood and frost
72 and does not require special soil conditions but prefers sandy and alkaline soils. As a
73 perennial grass it may live up to 10-15 years. Its industrial uses are well documented but there
74 are only limited data available on its natural element composition or requirement and
75 accumulation as well as tolerance to toxicity [8].

76 Heavy metals, such as cadmium, copper, lead, mercury, nickel and zinc are major
77 pollutants, particularly in areas with high anthropogenic pressure [9]. Szarvasi-1 energy grass
78 may be potentially applied in renewable energy production combined with phytoextraction or
79 phytostabilization. The aim of the present work was to assess the natural ability of Szarvasi-1
80 energy grass to accumulate or tolerate different heavy metals, Cd, Cu, Ni, Pb, Zn from
81 nutrient solution. Hydroponic culture was chosen for the experiments because it excludes the
82 different adsorption, mobility and retention characteristics of the metals in soil.

83

84 **2. Results**

85 2.1. Physiological responses to heavy metal treatments

86

87 The control and heavy metal containing nutrient solutions had very similar, slightly acidic pH
88 values which have been increased to slightly alkaline levels during cultivation of the plants
89 (Table 1). The extent of increase was smaller in case of Cd and Cu treatments.

90 Root and shoot growth was not affected, compared to the untreated control by Pb and
91 Zn applied in the nutrient solution in 10 μ M concentration for a month (Fig. 1). Ni and Cd
92 caused about 20 and 35% inhibition in the root and shoot growth, respectively. Cu had the

93 strongest effect on Szarvasi-1 decreasing the root and shoot dry mass by 90 and 75%,
94 respectively. When Cd and Cu were applied the relative dry matter content of the roots
95 increased with 76 and 138% whereas that of the shoots with 44 and 56%, respectively (Fig.
96 2).

97 The chlorophyll (Chl) concentration of the leaves changed most markedly under Cu
98 treatment that caused about 50% decrease leading to visible symptoms (Fig.3). Cd and Ni
99 caused a smaller but significant decrease compared to the control while the effect of Pb was
100 insignificant. The transpiration, measured as stomatal conductance for water vapour,
101 decreased by 79 and 91% in the plants treated with Cd and Cu, respectively (Fig. 4). Zn
102 increased the Chl concentration and stimulated the transpiration compared to the control,
103 although these changes were not significant.

104

105 2.2. Heavy metal concentration

106

107 Heavy metals applied in the treatments in 10 μM concentration were adsorbed by the roots in
108 different amounts (Fig. 5). Cd and Ni concentrations were very similar and the lowest among
109 the five metals. Zn concentration was twice as large while Cu concentration was almost 4
110 times larger than that of Cd and Ni. Pb was adsorbed in the highest amount ($280 \mu\text{mol g}^{-1} \text{DW}$
111 $= 1.35 \text{ mg kg}^{-1} \text{DW}$). Half of the roots of each plant were undertaken a washing procedure
112 (with $\text{CaSO}_4 + \text{Na}_2\text{EDTA}$) in order to remove the loosely bound part of the adsorbed metals.
113 The metal concentration was recalculated using the same dry mass data. The results revealed
114 that most of the Cd (79%) and Ni (93%) were not removable. In case of Zn only 41%
115 remained in the roots after washing. However, the root concentrations of the three metals
116 were statistically not different after the washing ($26\text{-}27 \mu\text{mol g}^{-1} \text{DW}$). In case of Cu and Pb

117 most of the adsorbed amount was removed in the washing procedure: only 8 and 1%
118 remained, respectively, resulting in the lowest concentrations.

119 In the shoot, Cd and Cu concentrations were similar while Ni and Pb were
120 significantly lower. Zn concentration was the highest reaching $4.71 \mu\text{mol g}^{-1}$ DW (300 mg kg^{-1}
121 DW) which is 0.03% of the shoot dry mass.

122

123 2.3. Essential metal concentration

124

125 The concentration of Fe was similar in the roots and shoots of control, Ni, Pb and Zn treated
126 plants (Fig. 6). However great difference was found between the Cd and Cu treated plants.

127 The latter two treatments caused a very high increase in the total Fe concentration of the

128 roots, 140 and 400 % by the Cd and Cu treatment, respectively, and the non-removable

129 fraction was still several times higher than in the control. The shoot concentration was

130 significantly changed (increased) only by Zn. Total Mn concentrations in the roots of Ni and

131 Pb treated plants were the same as in the control (Fig.7). Zn treatment reduced the Mn

132 concentration to one third of the control while Cd and Cu further reduced it to a minimal

133 level. The non-removable fraction of Mn accounted for 33-53% of the total amount. The

134 shoot contained Mn at a similar level in the control, Ni and Pb treatment while Zn slightly

135 reduced it. Cd and Cu decreased the shoot Mn to about half of the control.

136 Zn and Cu concentrations of the plants were compared in Figs. 8 and 9 when they

137 were applied at low (microelement) concentration. Only the roots of Cd and Ni treated plants

138 contained Zn in significantly lower concentration compared to the the control (Fig.8).

139 However, the non-removable fraction was different from the total only in the control. In the

140 shoot, only the Cu treated plants contained Zn at a lower level than the control. Cu

141 concentrations were almost identical to the control in the Ni, Pb and Zn treatment in both

142 roots and shoots while in the Cd treated plants it was 50% higher in the root and 20% lower in
143 the shoot (Fig.9).

144

145 **3. Discussion**

146

147 3.1. Heavy metal uptake

148

149 The metal content of shoot tissues depends on the uptake and translocation ability of root and
150 vascular tissues. In the hydroponic culture, the root system of Szarvasi-1 energy grass
151 adsorbed highly different amounts of heavy metals at slightly acidic to slightly alkaline pH in
152 the order: Pb>Cu> Zn> Cd≥Ni (Fig 5.). This finding is in agreement with previous work on
153 tall wheatgrass [10] and may be explained by different mechanisms. Cu, Zn and Ni are
154 essential transition metals required for normal growth in the order Zn>Cu> Ni and are readily
155 soluble in the applied experimental conditions. Their adsorbance may be driven by active
156 uptake. Cd is a nonessential heavy metal that is present in the nutrient solution in free divalent
157 ionic form [11]. However, after applying a washing procedure (CaSO₄ + Na₂EDTA solution)
158 in order to remove the portion deposited only to the apoplastic spaces, we found that Cd, Ni
159 and Zn were taken up by the roots in very similar amount, while Cu uptake was smaller. The
160 influx of transition metals is mediated by specific transporter proteins.

161 Zn uptake was first reported to be regulated by ZIP family genes in *Arabidopsis*
162 *thaliana* [12,13], however, the exact function of ZIPs is poorly known, yet [14]. In rice, a
163 Strategy II plant in Fe uptake, OsZIP1 and OsZIP3 seems to be important for Zn uptake from
164 soil [15,16]. In barley, Zn-DMA (deoxy mugineic acid – a phytosiderophore released by the
165 plant) is preferred over Zn²⁺ for uptake through roots [17]. In contrast, rice plants absorb less

166 Zn-DMA compared to Zn^{2+} [18]. Tall wheatgrass is a close relative to barley, thus DMA-
167 chelated Zn uptake can be predicted.

168 Cd may enter the root cells using different pathways provided by ZIP family
169 transporters, ZNT1 [19] and IRT1 which latter mediates Fe^{2+} uptake in non-graminaceous
170 plants [20] but was also found in rice [21]. In rice, Cd^{2+} uptake into the symplasm was shown
171 to be linked to Ca^{2+} transport, as accumulation of Cd is inhibited by La^{3+} and high Ca^{2+}
172 concentrations [22]. Wheat LCT1 (low-affinity cation transporter) was shown to have a role in
173 both Cd^{2+} and Ca^{2+} uptake [23]. In rice, OsNramp5 and OsNramp1 were reported as a root
174 plasma membrane transporter of Mn^{2+} and Cd^{2+} [24] and Fe^{2+} and Cd^{2+} [25], respectively.

175 The uptake and translocation of Cu is little known, it was found that P-type heavy
176 metal ATPases (HMAs) are involved [26,27]. Cu^+ transport into the cytosol is also mediated
177 by COPT family transporters in *A. thaliana* [28]. In graminaceous plants, the uptake of Cu
178 (and Zn) may be mediated by the release of phytosiderophores which (is increased under Fe
179 and Zn deficiency and) plays a distinct role in Fe acquisition [29]. ZIP2 and ZIP4 proteins are
180 also suggested to be transporting Cu^{2+} in *Arabidopsis* [30]. Cu^{2+} can form stable NA chelate
181 even under mild acidic conditions which complexes may have a role in the Cu translocation.
182 The uptake of Cu^{2+} -chelates cannot be excluded in Strategy-II plants, either. Gunawardana et
183 al [31] showed that Cu uptake is enhanced by the presence of histidine in the hydroponic
184 solution in ryegrass (*Lolium perenne*).

185 The uptake and translocation of Ni is poorly known, too. Ni may enter the cells in a
186 rather unspecific route through plasmalemma CNGCs (cyclic nucleotide gated channels) [32].
187 Nishida et al. [33] showed that AtIRT1, the primary Fe^{2+} uptake transporter in the root,
188 mediates Ni accumulation in *Arabidopsis thaliana*. But there is no evidence for a specific Ni
189 uptake in Strategy-II plants up to now.

190 Taking all these into account, ZIP-family transporters or chelation based strategies
191 (NA and DMA chelation) may be involved in the uptake of Cu, Ni and Zn. Cd uptake may
192 also interfere with that of Fe and Mn. Thus, regular disturbances in the essential transition
193 metal uptake and translocation in heavy metal treated Szarvasi-1 energy grass can be
194 explained as complex interference in these systems.

195 The various uptake mechanisms described above show that probably it is not the way
196 of influx that matters as it does not provide explanation for the higher adsorption and lower
197 uptake of Cu compared to the other transition metals in Szarvasi-1 energy grass. The
198 formation of Fe-plaque in the roots of Cu treated plants may account for the retention of Cu
199 and also Fe on the roots [34]. Such an unspecific mechanism may be predicted in the case of
200 Cd and Pb, too. In Cd treated plants this is underlined by the accumulation of Fe and Cu in the
201 root apoplast. Pb is a nonessential heavy metal that may produce relatively insoluble
202 precipitates with the constituents of the nutrient solution in sulphate and phosphate (or
203 chloride) form on the root surface [35]. Soluble Pb concentration can be increased with
204 complexing agents like EDTA or citrate [8] which serves as the basis for Pb mobilization in
205 polluted soils during “induced phytoextraction” [36,37,38]. Applying the (CaSO₄ +
206 Na₂EDTA) washing procedure, we found that Pb taken up by the roots was almost negligible.
207 Pb uptake may be mediated either by CNGC [32] or P-type ATPase transporters [39].
208 However, as most of this metal is removable from the root apoplast its uptake may occur
209 through more unspecific routes, too. The ionic radius of Pb²⁺ is much larger compared to the
210 other metal ions tested, thus it may be assumed to surge into stelar tissues through internal
211 wounding by lateral root formation [40] or may be taken up by endocytosis [41].

212

213 3.2. Heavy metal translocation

214

215 The shoot metal concentration in our study increased in the order: Pb<Ni<Cu~Cd<Zn (Fig.
216 5). This is in contradiction with the finding of Yang et al. [10] even though data were
217 compared after recalculation based on their figures. They found that the accumulation
218 depended on the applied concentration and it increased in the order Pb<Cd<Cu<Ni at 0.5 mM
219 metal dose. The authors applied chloride form of the metals and much higher concentrations
220 than the present work. Although, the accumulation order was different, Pb accumulation in
221 the shoot was the smallest, too. In our previous work, it was found that Pb accumulation in the
222 shoot may even be lower if the Fe-chelator applied in the nutrient solution is EDTA while
223 citrate may be effective in increasing shoot Pb accumulation at higher Pb levels in the
224 medium [8]. The low Pb uptake and accumulation was shown also by the low PC values
225 (Table 2). Once loaded into the xylem, Pb may be transported in Pb²⁺-citrate complex form as
226 it was suggested previously [42].

227 Nickel transport was so low that hardly exceeded that of Pb in Szarvasi-1 energy grass
228 (Fig. 5) whereas in Yang et al. [10] it has the highest concentration in the shoot. However, the
229 shoot Ni concentration we found and its translocation (TF=0.020, Table 2) is very similar to
230 that in Chen and Wong [43] (TF=0.03-0.06 depending on soil Ni concentration at pH 8,
231 calculated from the published data) who worked with tall wheatgrass grown in soil.
232 Concerning the mechanism of its translocation, histidine was shown to interfere with the
233 xylem loading of Ni²⁺ [44]. Its transport in the xylem sap was suggested to occur in chelated
234 form [45].

235 The very low accumulation of Pb and Ni in the shoot implies that their translocation in
236 Szarvasi-1 energy grass is driven only by transpiration and that there is no metabolic demand
237 for Ni, either.

238 Cu and Cd concentrations in the shoot were very similar in the treated plants and their
239 value was also similar to Fe concentration in all treatments. Xylem loading of Cu and Cd may

240 occur by active efflux through P-type ATP-ases [28,46]. Gunawardana et al. [31] also
241 reported a similar behaviour of Cu and Cd so that additional citric acid enhanced the
242 translocation of both metals. Fe is translocated as Fe³⁺-citrate complexes in the xylem sap
243 [47]. Curie et al. [48] showed that Cu-NA complex is completely stable at the pH of xylem
244 sap (pH 5–6) and Cu is transported to the shoot in NA-chelated form. The synthesis of
245 chelators may increase upon Cu excess [49] while the long-distance translocation of Cd,
246 which does not form chelates under *in vivo* conditions, may depend on the availability of
247 other elements [50], and is less dependent on the presence of chelators in the xylem sap. Cu
248 concentration increased 9-fold in the shoot of treated plants compared to the control but this
249 increase came together with severe toxicity symptoms discussed below. This finding is in
250 agreement with previous work [10]. Despite the similar shoot concentrations of Cu, Cd and
251 Fe, both Cu and Cd translocation was much higher than that of Fe in the treated plants which
252 latter was highly retarded by the treatments compared to the untreated control (Table 2). TF
253 values for Cu and Pb were very similar and also higher than those of Ni and Cd. However, the
254 PC of Cu and Cd was not considerably higher than that of Pb, and Ni and was the same as that
255 of Fe. Moreover, PC values of Pb and Ni were the lowest and also much lower than that of Fe
256 in the same plants. These findings show that Szarvasi-1 energy grass is not an efficient
257 accumulator of Cd, Cu, Ni and Pb. This is not the case for Zn.

258 Szarvasi-1 energy grass proved to be very efficient in accumulating Zn in the shoot.
259 Vetiver grass (*Chrysopogon zizanioides*) accumulated 6.2 µmol g⁻¹ Zn in the shoot (as
260 compared to 4.7 µmol g⁻¹ in Szarvasi-1 energy grass) but in that case much higher Zn
261 concentrations were measured in the soil solution [37]. Wheat genotypes are different
262 concerning their ability to take up and transport Zn. Zn efficient genotypes release more
263 phytosiderophores which correlates with higher shoot concentrations [51]. Furthermore,
264 Hacisalihoglu et al. [52] identified high and low affinity Zn transport systems in wheat roots,

265 while Zn translocation was shown to be very efficient resulting in balanced concentration in
266 the shoots of radiolabelled plants [53]. Zn-NA transporters or Zn-DMA transporters involved
267 in Zn translocation have not been identified, yet [16] but Zn-NA complexes were shown to
268 exist in the phloem sap of rice [54]. Ishimaru et al. [15] suggested that Zn deficiency induces
269 DMA synthesis in barley shoots, while both Zn and Fe deficiency induce MA synthesis and
270 secretion in barley roots. These data indicate that Strategy-II plants may efficiently scavenge
271 Zn from the soil. Although Cd and Zn are chemically very similar [55], they behave
272 differently. Cd showed one of the lowest while Zn had the highest TF as compared to the
273 other metals in the treatments while the PC value for Zn was 6.5 times higher than that of Cd
274 and twice as high even that that of Fe (Table 2). This indicates that Szarvasi-1 energy grass is
275 indeed very efficient in Zn accumulation.

276

277 3.3. Physiological dysfunctions under heavy metal treatments

278

279 The increase in the pH of the nutrient solution under all treatments can be explained by the
280 original habitat preference of tall wheatgrass: it grows in alkaline soils with pH 6-10 (Table
281 1). The pH increase during cultivation under Cd and Cu treatments was moderate which
282 implies a disturbed metabolism by these metals. This was confirmed by a serious growth
283 inhibition and higher dry matter content whereas neither of the other heavy metals caused any
284 disturbance in these parameters compared to the control except for Ni treatment which led to a
285 slight growth inhibition in the root and shoot (Figs 1, 2).

286 Interestingly, Pb had no significant effect on any physiological parameters measured
287 in this work. This may have been due to its very low concentration in the root symplast and in
288 the shoot.

289 Transition metals Cd, Cu and Ni reduced Chl concentration in Szarvasi-1 energy grass
290 (Fig. 3). Cd is known to decrease Chl concentration in Strategy-I plants by decreasing the
291 citrate transporter FRD3 expression in root xylem parenchyma which leads to decreased Fe
292 translocation [56]. Stomatal conductance also decreased under Cd treatment (Fig. 4). Cd is
293 known to interact with Ca metabolism resulting disturbed signalling processes. Thus, the
294 presence of Cd in guard cells leads to stomatal closure [57]. Decreased transpiration rate
295 enhances the inhibition of growth as well as metal translocation (see the decrease in the shoot
296 concentration of Cu and Mn, Figs. 7 and 9).

297 Ni reduced Chl concentration similar to Cd but its effect on growth was much less
298 pronounced. As it did not modify the uptake and translocation of essential elements and the
299 transpiration, either, it may have reduced only the photosynthetic performance of plants
300 through inhibition of Chl synthesis.

301 Cu decreased the Chl concentration in the plants to the highest extent which is
302 combined with a significant inhibition in the stomatal conductance. It decreased the shoot Zn
303 and Mn concentration compared to the control (Figs. 7, 8). Although the TF for Fe under Cu
304 treatment decreased severely (Table 2), Cu did not significantly interfere with Fe
305 concentration in the shoot (Fig. 6), the inhibition of Chl synthesis may have been coupled to
306 the inhibited development of the photosynthetic apparatus due to Mn deficiency and a
307 superimposed oxidative stress due to high Cu and low Zn concentration [58]. The lower Chl
308 concentration may have resulted in lower photosynthetic performance that required a lower
309 gas exchange rate leading to stomatal closure.

310 Zn slightly enhanced the Chl accumulation as well as transpiration (Figs. 3, 4). The
311 uptake of Zn may interact with Fe uptake resulting in higher non-removable Fe concentration
312 in the root and also higher shoot Fe concentration compared to the control (Fig. 6) that may
313 explain the positive effect on the mentioned parameters. Zn transport in the xylem is most

314 probably independent from that of Fe but there is a clear stimulation of Fe translocation along
315 with an inhibition of Mn translocation (Fig. 7) under Zn treatment. In fact, the PC calculated
316 for Fe under Zn treatment was higher than in the control (Table 2). Zn stress is known to
317 inhibit the growth and leaf expansion and may also lead to oxidative stress [59]. However,
318 these effects were not observed in Szarvasi-1 energy grass that implies a very efficient
319 detoxification mechanism in this plant.

320

321 **4. Conclusions**

322

323 Szarvasi-1 energy grass, an energy crop potentially applicable in phytoremediation was
324 shown to be sensitive to high external concentrations of Cu and Cd and fairly tolerant to Ni
325 and Pb. Cu and Cd toxicity leading to the inhibition of growth, transpiration and Chl
326 biosynthesis can be attributed to the uptake and translocation of these metals which in turn
327 causes imbalance in microelement homeostasis and most probably oxidative stress. Fe plaque
328 formation suggested by Fe and Cu accumulation in the roots may also explain the negative
329 effects. Tolerance to Ni and Pb can be explained by the very low concentration of these
330 metals in the shoot due to the unspecific nature of their uptake and transport. Szarvasi-1
331 showed high rates of Zn translocation to the shoot as compared to the other metals that was
332 combined with high tolerance. Exposure to high concentration of Zn even resulted in a slight
333 stimulation of growth, transpiration and Chl biosynthesis in paralel with its accummulation in
334 the shoot. Based on these observations the plant is eligible for phytostabilization of Ni-Pb-Zn
335 contaminated soils that has to be confirmed with soil-based experiments.

336

337 **5. Materials and Methods**

338

339 5.1. Plant material and treatments

340

341 The seeds of the tall wheatgrass cultivar Szarvasi-1 energy grass (*Elymus elongatus* subsp.
342 *ponticus* (Podp.) Melderis cv. Szarvasi-1 (syn. *Agropyron elongatum*, *Elytrigia elongata*,
343 Csete et al., 2011), developed for industrial purposes were germinated for five days on wet
344 filter papers in Petri dishes at room temperature and sunlight. Ten seedlings with 2-5 cm long
345 roots were placed on a 2 cm wide strip of sponge-rubber, rolled up and fastened in a
346 polystyrene ring and they were transferred to plastic containers. Each container was filled up
347 with 0.7 dm³ modified quarter strength Hoagland nutrient solution of the following
348 composition: 1.25 mM KNO₃; 1.25 mM Ca(NO₃)₂; 0.5 mM MgSO₄; 0.25 mM KH₂PO₄; 11.6
349 μM H₃BO₃; 4.5 μM MnCl₂.4H₂O; 0.19 μM ZnSO₄.7H₂O; 0.12 μM Na₂MoO₄.2H₂O; 0.08 μM
350 CuSO₄.5H₂O and 10 μM Fe(III)-citrate-hydrate. Metals were added as Cd(NO₃)₂,
351 CuSO₄.5H₂O, NiSO₄, ZnSO₄.7H₂O, Pb(NO₃)₂ in 0 or 10 μM concentration to the nutrient
352 solution, separately. Fresh solutions were used for cultivation without buffering or other pH
353 adjustment. The plants modified the original pH of the solutions at the same level during the
354 period between solution changes (Table 1).

355 Plants were grown in a climate controlled growth chamber at 20/25 °C, at 75% relative
356 humidity and 150 μmol m⁻² s⁻¹ PPFD with 10/14h dark/light period. The nutrient solution was
357 continuously aerated and replaced fresh solution twice a week.

358 Three parallel pots, each containing 10 plants were applied for a treatment group
359 ending up in 18 pots with the untreated control and the whole experiment was carried out
360 twice. Physiological parameters were measured and the plants were harvested 37 days after
361 germination.

362

363 5.2. Mass measurements

364

365 The roots of the 10 plants grown in a single pot were separated into two portions. The roots of
366 the first 5 plants were centrifuged between filter papers at 300 g to remove traces of nutrient
367 solution before drying but no other treatment was applied. The other 5 roots were rinsed with
368 0.5 mM CaSO₄ solution and then transferred to 200 ml 0.5 mM CaSO₄ solution containing 10
369 mM Na₂EDTA (pH 4.05) and were shaken for 1.5 h at 125 rpm [11]. After rinsing again with
370 CaSO₄ the roots were centrifuged between filter papers at 300 g. The filtered roots were
371 weighed. Dry mass was determined after drying at 80 °C. Final data are extrapolated to one
372 single plant.

373

374 5.3. Element analysis

375

376 Measurements were made with three parallel samples (each containing 5 dried plants) after
377 acidic digestion. 5-10 ml cHNO₃ was added to each gram of the samples for overnight
378 incubation. Then the samples were pre-digested for 30 min at 60 °C. Finally, 2-3 ml H₂O₂ (30
379 m/m%) was added for a 90 min boiling at 120 °C. The solutions were filled up to 10-50 ml,
380 homogenised and filtered through MN 640W filter paper. The element content of the filtrate
381 was determined by ICP-MS. Data were converted from ppm to μmol g⁻¹ units in order to
382 ensure better comparison between treatments.

383

384 5.4. Chlorophyll concentration

385

386 The measurements were made with the first fully developed leaves. The Chl concentration
387 was determined photometrically (Shimadzu UV-2101PC) from 80% acetone extracts using

388 the equations of Porra et al. [60]. Each measurement was carried out on three individual
389 plants in each treatment group.

390

391 5.5. Stomatal conductance

392

393 Stomatal conductance was measured with a porometer (DELTA-T Devices Ltd.) on the
394 abaxial epidermis of the middle sections of the youngest, fully developed leaves.
395 Transpiration was calculated as $\text{mmol H}_2\text{O m}^{-2} \text{ s}^{-1}$. Each measurement was carried out three
396 times on three individual plants in each treatment group.

397

398 5.6. Definition of indices

399

400 Translocation factor (TF) and phytoextraction capacity (PC) was defined after Vashegyi et al.
401 [8] with some modifications. Translocation factor of $\text{Me}_i = \text{shoot Me}_i \text{ concentration (mol g}^{-1}\text{)}$
402 $/ \text{Me}_i \text{ concentration in the washed roots (mol g}^{-1}\text{)}$. Phytoextraction capacity of $\text{Me}_i = \text{shoot}$
403 $\text{total Me}_i \text{ content (g) * 100 / total amount of Me}_i \text{ supplied to the nutrient solution during the}$
404 $\text{entire growth period (g)}$

405

406 5.7. Statistics

407

408 Basic statistical analysis was carried out with one-way ANOVA and Tukey-Kramer multiple
409 comparisons test ($p < 0.05$) using Statistica 2000 (Statsoft) and InStat 3.0 (GraphPad)
410 softwares.

411

412

413 **Acknowledgements**

414

415 We would like to thank Zsuzsa Ostorics for her technical assistance.

416

417

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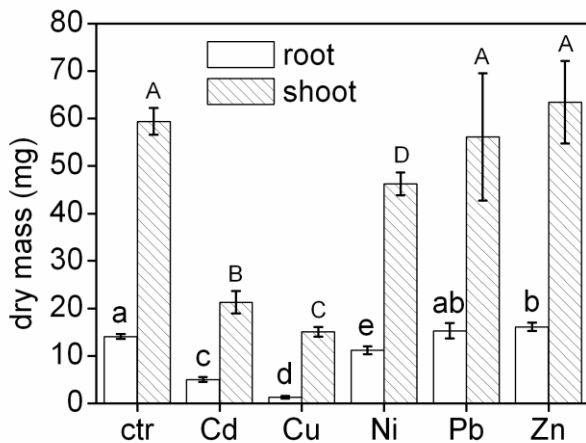
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582 Captions to Figures

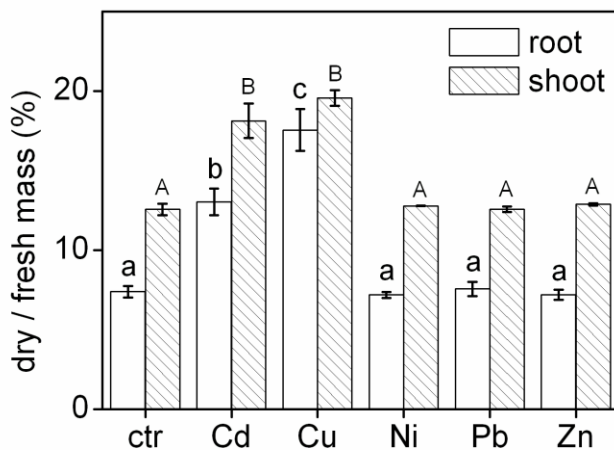
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584 Figure 1 Root and shoot dry mass of 37 day-old Szarvasi-1 energy grass grown in nutrient
585 solutions amended with different metals (Cd, Cu, Ni, Pb, Zn) in 0 or 10 μ M concentration.
586 (Data are shown as mean \pm SD, n=6, significant differences between data are indicated with
587 different letters, P<0.05)



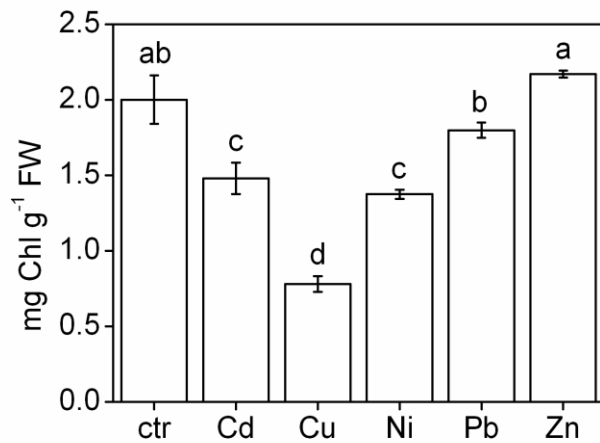
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589 Figure 2 Dry matter content in the roots and shoots of 37 day-old Szarvasi-1 energy grass
590 grown in nutrient solutions amended with different metals (Cd, Cu, Ni, Pb, Zn) in 0 or 10 μ M
591 concentration. (Data are shown as mean \pm SD, n=6, significant differences between data are
592 indicated with different letters, P<0.05)

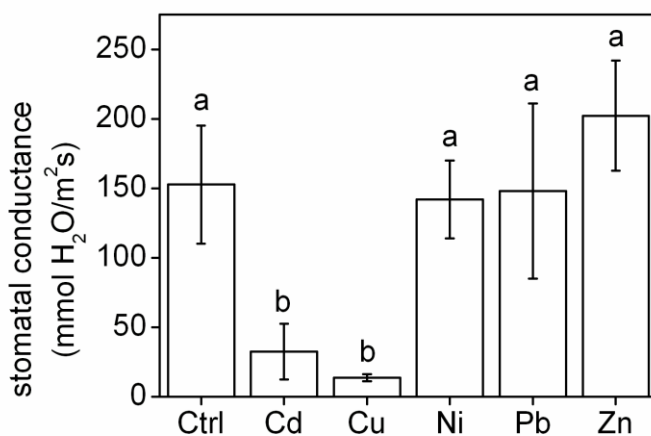


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594 Figure 3 Chlorophyll concentration in the leaves of 37 day-old Szarvasi-1 energy grass grown
595 in nutrient solutions amended with different metals (Cd, Cu, Ni, Pb, Zn) in 0 or 10 μM
596 concentration. (Data are shown as mean \pm SD, n=6, significant differences between data are
597 idicated with different letters, P<0.05)

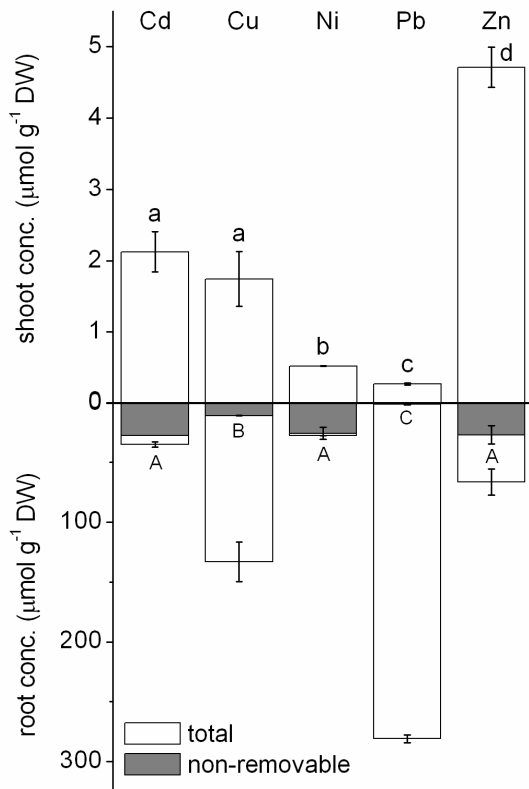


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599 Figure 4 Transpiration of the leaves of 37 day-old Szarvasi-1 energy grass grown in nutrient
600 solutions amended with different metals (Cd, Cu, Ni, Pb, Zn) in 0 or 10 μM concentration,
601 measured as the stomatal conductance for water vapour. (Data are shown as mean \pm SD, n=9,
602 significant differences between data are idicated with different letters, P<0.05)



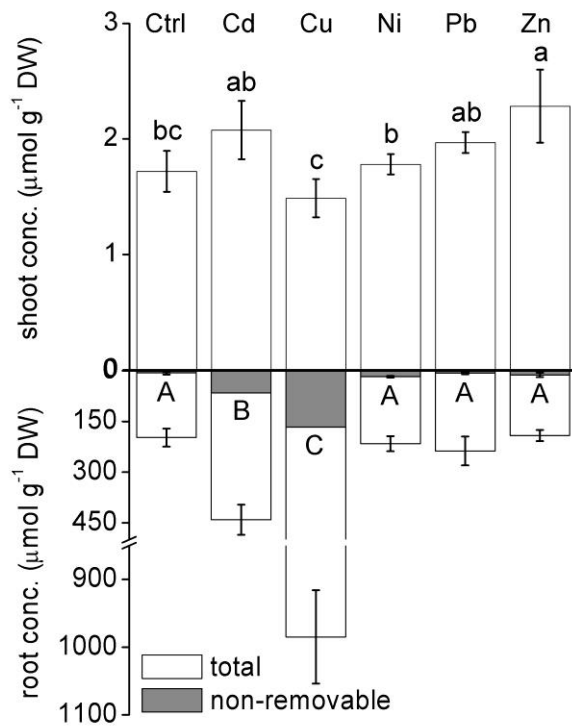
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605 Figure 5 Heavy metal concentration in the roots and shoots of 37 day-old Szarvasi-1 energy
 606 grass grown in nutrient solutions amended with the different metals (Cd, Cu, Ni, Pb, Zn) in 0
 607 or 10 μM concentration. Shaded parts of the coloumns show the concentration of metals non-
 608 removable by washing with $\text{CaSO}_4+\text{Na}_2\text{EDTA}$ solution. (Data are shown as mean \pm SD, n=6,
 609 significant differences between data are idicated with different letters, $P<0.05$)



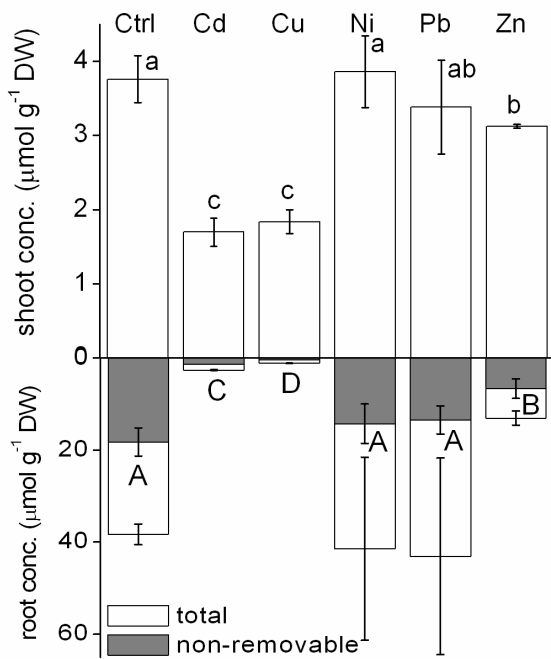
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618 Figure 6 Fe concentration in the roots and shoots of 37 day-old Szarvasi-1 energy grass grown
 619 in nutrient solutions amended with different metals (Cd, Cu, Ni, Pb, Zn) in 0 or 10 μM
 620 concentration. Shaded parts of the coloumns show the concentration of Fe non-removable by
 621 washing with $\text{CaSO}_4+\text{Na}_2\text{EDTA}$ solution. (Data are shown as mean \pm SD, n=6, significant
 622 differences between data are idicated with different letters, $P<0.05$)



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631 Figure 7 Mn concentration in the roots and shoots of 37 day-old Szarvasi-1 energy grass
 632 grown in nutrient solutions amended with different metals (Cd, Cu, Ni, Pb, Zn) in 0 or 10 μM
 633 concentration. Shaded parts of the columns show the concentration of Mn non-removable by
 634 washing with $\text{CaSO}_4+\text{Na}_2\text{EDTA}$ solution. (Data are shown as mean \pm SD, n=6, significant
 635 differences between data are indicated with different letters, $P<0.05$)



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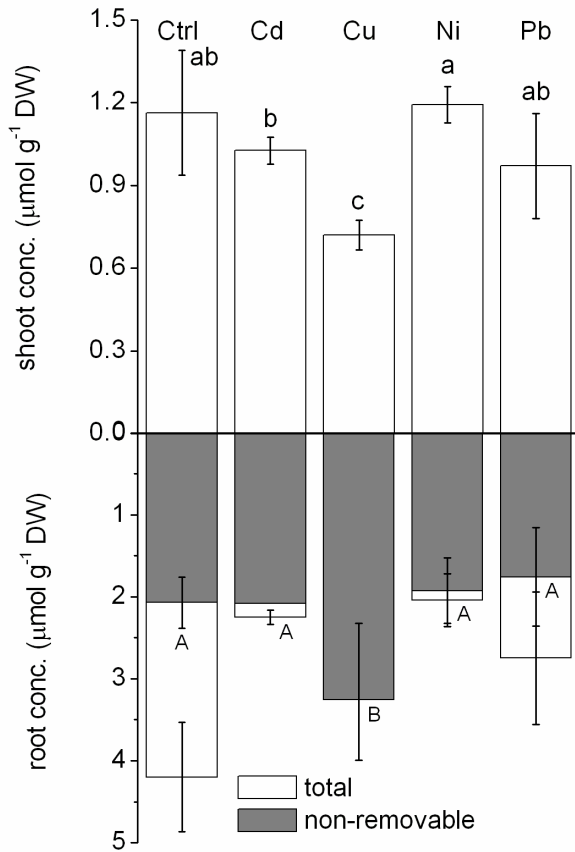
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643 Figure 8 Zn concentration in the roots and shoots of 37 day-old Szarvasi-1 energy grass
 644 grown in nutrient solutions amended with different metals (Cd, Cu, Ni, Pb) in 0 or 10 μM
 645 concentration. Shaded parts of the columns show the concentration of Zn non-removable by
 646 washing with $\text{CaSO}_4+\text{Na}_2\text{EDTA}$ solution. (Data are shown as mean \pm SD, n=6, significant
 647 differences between data are indicated with different letters, $P<0.05$)



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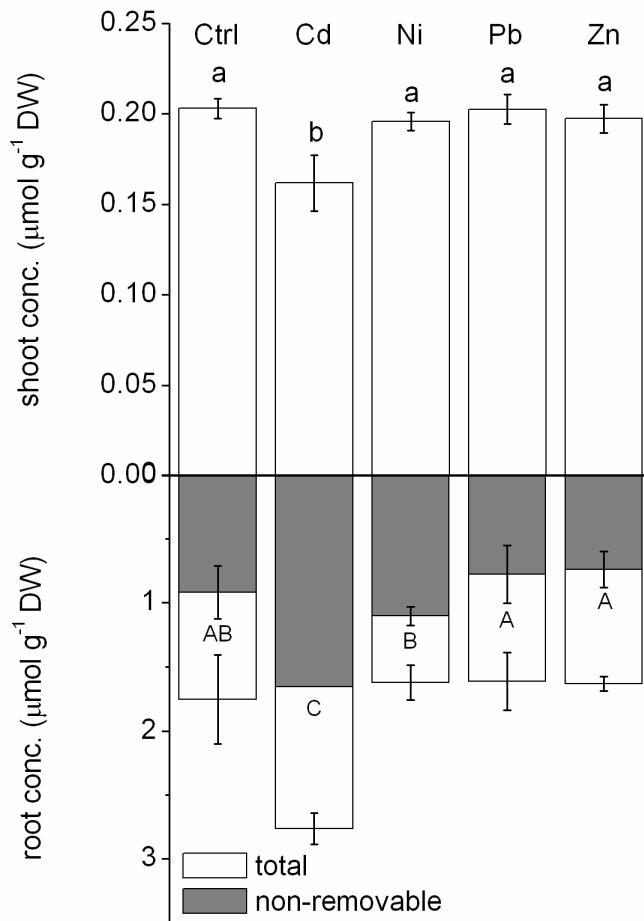
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656 Figure 9 Cu concentration in the roots and shoots of 37 day-old Szarvasi-1 energy grass
 657 grown in nutrient solutions amended with different metals (Cd, Ni, Pb, Zn) in 0 or 10 μM
 658 concentration. Shaded parts of the columns show the concentration of Cu non-removable by
 659 washing with $\text{CaSO}_4+\text{Na}_2\text{EDTA}$ solution. (Data are shown as mean \pm SD, n=6, significant
 660 differences between data are indicated with different letters, $P<0.05$)



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667 Table 1. pH values of the nutrient solutions amended with different heavy metals (Cd, Cu, Ni,
668 Pb, Zn) in 0 (ctr) or 10 μ M concentration at preparation (Day 0) and after 4 days of plant
669 growth (Day 4) in unbuffered, aerated hydroponic culture of one month-old Szarvasi-1 energy
670 grass. (Data are presented as mean \pm SD, n=6, significant differences between data are indicated
671 with different letters, P<0.05))

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673	treatment	Day 0	Day 4
674	ctr	4,70	7,66 \pm 0,05 a
675	Cd	4,67	6,73 \pm 0,01 b
676	Cu	4,70	6,11 \pm 0,08 c
677	Ni	4,60	7,42 \pm 0,12 d
678	Pb	4,76	7,297 \pm 0,20 d
679	Zn	4,78	7,72 \pm 0,05 a

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691 Table 2. Translocation factor [TF = shoot Mei concentration (mol g⁻¹) / Mei concentration of
 692 the washed roots (mol g⁻¹)] and phytoextraction capacity [PC = shoot total Mei content (g)
 693 *100 / Mei supplied to the nutrient solution during the whole growth period (g)] of Mei and
 694 Fe in Szarvasi-1 energy grass grown in nutrient solutions amended with different metals
 695 (Mei) in 0 (ctr) and 10 µM concentration. (The concentration of Fe was also 10 µM in all
 696 treatments and the untreated control.)

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698 treatment (Mei)	TF (Mei)	TF (Fe)	PC (Mei)	PC (Fe)
699 Cd	0,077	0.031	0.071	0.069
700 Cu	0.159	0.009	0.041	0.035
701 Ni	0.020	0.102	0.038	0.129
702 Pb	0.160	0.281	0.024	0.173
703 Zn	0.174	0.188	0.467	0.226
704 Fe (ctr)	-	0.265	-	0.159