



**Biodiversity in managed and unmanaged forests:
compromises in data selection for meta-analysis do not
mean compromised results. Reply to Halme et al.**

Journal:	<i>Conservation Biology</i>
Manuscript ID:	Draft
Wiley - Manuscript Category:	Comment
Date Submitted by the Author:	
Complete List of Authors:	Paillet, Yoan; Cemagref, UR EFNO
Abstract:	

1 **Biodiversity in managed and unmanaged forests: compromises in data**
2 **selection for meta-analysis do not mean compromised results. Reply to Halme**
3 **et al.**

4 Yoan Paillet^{1,2}, Laurent Bergès^{1,*}, Joakim Hjältén³, Péter Ódor⁴, Catherine Avon¹, Markus
5 Bernhardt-Römermann⁵, Rienk-Jan Bijlsma⁶, Luc De Bruyn^{7,8}, Marc Fuhr², Ulf Grandin⁹,
6 Robert Kanka¹⁰, Lars Lundin⁹, Sandra Luque², Tibor Magura¹¹, Silvia Matesanz¹², Ilona
7 Mészáros¹³, M.-Teresa Sebastià^{14,15}, Wolfgang Schmidt⁵, Tibor Standovár⁴, Béla
8 Tóthmérész¹⁶, Anneli Uotila¹⁷, Fernando Valladares¹², Kai Vellak¹⁸, Risto Virtanen¹⁹

9 *corresponding author, mailto: laurent.berges@cemagref.fr

10
11 ¹ Cemagref, UR EFNO, Domaine des Barres, F-45290 Nogent-sur-Vernisson, France

12 ² Cemagref, UR EMGR, 2 rue de la Papeterie BP 76, F-38402 Saint-Martin-d'Hères, France

13 ³ Department of Wildlife, Fish and Environmental Science, Swedish University of Agricultural Sciences, SE-901 83 Umeå,
14 Sweden

15 ⁴ Department of Plant Taxonomy and Ecology, Eötvös University, Pázmány P. stny. 1/C., H-1117 Budapest, Hungary

16 ⁵ Department Silviculture and Forest Ecology of the Temperate Zones, Georg-August-University Göttingen, Büsgenweg 1, D-
17 37077 Göttingen, Germany

18 ⁶ Alterra Wageningen UR, Centre for Ecosystem Studies, P.O. Box 47, NL-6700 AA Wageningen, The Netherlands

19 ⁷ Research Institute for Nature and Forest, Kliniekstraat 25, B-1070 Brussels, Belgium

20 ⁸ Evolutionary Ecology, Department of Biology, University of Antwerp, Groenenborgerlaan 171, B-2020 Antwerpen, Belgium

21 ⁹ Swedish Univ Agr Sci, Department of Aquatic Sciences and Assessment, Box 7050, SE-75007 Uppsala, Sweden

22 ¹⁰ Institute of Landscape Ecology, Slovak Academy of Sciences, Stefanikova Str. 3, SK-814 99 Bratislava, Slovakia

23 ¹¹ Hortobágy National Park Directorate, P.O. Box 216, H-4002 Debrecen, Hungary

24 ¹² Instituto de Recursos Naturales, CSIC IRN-CCMA-CSIC, Serrano 115, E-28006 Madrid, Spain

25 ¹³ Department of Botany, University of Debrecen, P.O. Box 71, H-4010 Debrecen, Hungary

26 ¹⁴ Forest Technology Centre of Catalonia, Pujada del Seminari s/n, E-25280 Solsona, Spain

27 ¹⁵ Agronomical Engineering School, University of Lleida, Av. Rovira Roure 191, E-25198 Lleida, Spain

28 ¹⁶ Ecological Institute, Debrecen University, P.O. Box 71, H-4010 Debrecen, Hungary

29 ¹⁷ Faculty of Forestry, University of Joensuu, P.O. Box 111, FIN-80101 Joensuu, Finland

30 ¹⁸ Institute of Ecology and Earth Sciences, University of Tartu, Lai Str., 40 Tartu EE-51005, Estonia

31 ¹⁹ Department of Biology, University of Oulu, P.O. Box 3000, FIN-90014 Oulu, Finland

32
33 Running head: Reply to Halme et al.

34
35 Word count: 2129 words

36

37 **Introduction**

38 Meta-analysis (hereafter MA) is a powerful tool to assess general trends and quantitatively
39 synthesize the results of independent studies. However, this procedure has received
40 criticisms, particularly when it has been applied to ecological and conservation biology
41 studies.

42 In an attempt to provide a state of the art of the effect of forest management on biodiversity,
43 we performed a MA comparing the species richness of managed and unmanaged forests in
44 Europe (Paillet et al. 2010). We intended to review and analyze the recent publications
45 regarding the biodiversity of management of forests. Thus, the opening sentence of Halme et
46 al. (this issue) goes against the goal and the basic philosophy of our paper. Indeed, Paillet et
47 al. (2010, p. 103) provided a balanced view of the contrasting opinions.

48 Our MA provides basic ecological knowledge needed for conservation and ecologically
49 sustainable forestry. In this paper, we showed that forest management has a negative effect
50 on the biodiversity of forest dwelling species. Because we were aware of the limitations of
51 our MA, we used caution when discussing the results considering that: (i) the effect is
52 strongly heterogeneous between different taxa; (ii) there is a *trend* for recovery of biodiversity
53 once management has been abandoned; (iii) no strong conclusion on the effect of different
54 management types could be drawn from our data due to low replication number. The obvious
55 main conclusion of this paper was that research on the subject in Europe was scarce and
56 that more controlled studies may help answer the questions raised.

57 However, Halme et al. surprisingly overlook the fact that confounding effects and MA
58 limitations were largely discussed in our paper (p. 109-110). Further, they have claimed that
59 our data selection has four major flaws that compromise our conclusions: independence of
60 observations, distribution of the taxonomic groups regarding time since abandonment and
61 management intensity, taxonomic generalizations and criteria used for inclusion of papers.

62

63 **Independence of observations**

64 We share Halme et al.'s concern on proper replication in scientific studies. However, the
65 pseudoreplication issue is much more complex than Halme et al. indicate. Specifically in the
66 case of large scale field experiments, the question on what comprises a replicate has been
67 intensely debated (e.g. Oksanen 2001, 2004; Underwood 1997). In addition, it has been
68 argued that the core ideas behind pseudoreplication are based on a misunderstanding of
69 statistical independence, the nature of control groups in science, and contexts of statistical
70 inference (Koehnle & Schank 2009; Schank & Koehnle 2009 but see also Hurlbert 2009).
71 Unfortunately, this issue is too complicated to be explored in such a short reply but this
72 underscores the complexity of the pseudoreplication question.

73 Although Hurlbert's (1984) paper pinpointed a very important problem in ecology, because of
74 the complexity of this issue, it cannot and should not be used as a universal criterion for
75 accepting or rejecting experimental research; all research must be judged on its own merits.
76 For example, spatiotemporal proximity does not automatically lead to statistical dependence
77 and certainly not in a way that prohibits appropriate statistical inferences (Schank & Koehnle
78 2009). In addition, there are many other important methodological and statistical issues to
79 consider when evaluating the quality of a piece of research. Contrary to manipulative
80 experiments, we should accept that background variations cannot be fully controlled in
81 mensurative experiments (Hurlbert 1984), and it is often impossible to spatially replicate on a
82 large number of different sites. As clearly stated in our article, the surface area of
83 unmanaged forests is very limited in Europe. Comparatively, the number of managed forest
84 stands is much higher. Thus, it is nearly impossible to control for important factors like site
85 conditions, patch size, landscape context, soil, stand age, tree species composition and land
86 use history. Avoiding pseudoreplication is a desired prerequisite of many experiments, but it
87 is practically impossible to fulfil when investigating unmanaged forests in Europe. Thus, we
88 disagree with Halme et al. that some papers should have been excluded from our meta-
89 analysis due to supposed pseudoreplication problems. These papers have all been
90 published in peer-reviewed journals and have thus been judged to have scientific merit. If we
91 were to subjectively exclude from our analyses the papers we consider flawed (for

92 methodical, statistical or other reasons), our objectivity could be questioned. We therefore
93 chose to include these studies in our analyses.

94 Halme et al. also questioned the independence of observations, but in the MA process,
95 comparing a single control to several experimental groups is generally accepted (Gurevitch &
96 Hedges 2001). These cases finally represent 22% of the total number of comparisons (26 out
97 of 120 in our dataset). Sampling dependence in multiple-treatment studies can be solved in
98 three ways: (i) by using the unmanaged forest stands just once and randomly choosing one
99 managed forest types and leaving out the other types; (ii) by mixing all the managed plots in
100 one; (iii) by using a meta-analysis model with study as a random effect, which controls for
101 this type of dependence (Gurevitch & Hedges 2001).

102

103 **Distributions of co-variables**

104 The added-value of a MA relies on its ability to test the relationship between effect size and
105 factors that were not testable in the individual studies (Gurevitch & Hedges 2001). The
106 distribution across different taxa and covariables is definitely unbalanced, as emphasized by
107 Table 1 and Table 1 p. 104 in Paillet et al. (2010). However, we do not share the opinion that
108 the general trend observed between effect size and Time Since Abandonment (TSA) effect
109 was only an artefact of the unbalanced distribution between vascular plants studies on the
110 one hand and fungi and carabid beetles on the other. Figure 2, p. 107 in Paillet et al. (2010)
111 clearly shows that there are many negative effect sizes around 50 years and positive ones
112 around 100 years; this partly counterbalances the distribution at the extreme end of the TSA
113 gradient. Moreover, Halme et al. do not mention that analyses separated by taxa almost
114 always provided negative slopes, except for bryophytes and birds (see Table 3, p. 107).
115 Finally, even if the effect of TSA was significant only for carabids, saproxylic beetles and
116 fungi, most of the negative slopes for taxa have much higher value than the slope for all
117 groups.

118 Halme et al. criticised the extrapolation of the regression equation because the TSA values
119 for which effect size equalled zero were outside the range of observed TSA for carabids and

120 fungi. However, 43 years is very close to the minimum TSA for fungi (50 years). More
121 generally, we trust the readers of our article to only consider the threshold values we
122 provided as indicative since we never claimed that these constitute absolute references for
123 forest management policy.

124 The example of management intensity, far from nullifying our results, actually confirms and
125 strengthens the conclusion clearly presented in the abstract (p. 102), results (p. 107) and
126 discussion (p. 109) sections: low replication number and poor information on management
127 methods are not sufficient to conclude on the effect of management type. More generally,
128 MA methods are still under development: the test of interaction between factors is not yet
129 implemented in statistical software and this is a challenging issue. However, low replication
130 number would prevent us from testing interactions in a robust way.

131

132 **Taxonomic generalisations**

133 Concerning fungi and saproxylic beetles, our systematic research identified the studies
134 currently available. Certainly, the fact that the fungi kingdom is mainly represented by taxa
135 dependent on deadwood should have been mentioned in the tables, but this is clearly stated
136 in the discussion. Concerning bark beetles, although we agree that some are early-
137 successional species favoured by forestry, e.g. clear-felling, the majority of them are not, and
138 many are confined to old-growth forests. If we analyse the two groups separately, we obtain
139 the same trend and can consequently draw the same conclusions: the mean effect size was
140 negative and significant for bark beetles ($d_{+} = -0.76$, bootstrap CI = -1.21 to -0.35, $n=6$) and
141 negative but marginally significant for the other saproxylic beetles ($d_{+} = -0.65$, bootstrap CI = -
142 1.41 to -0.01, $n=11$). Contrary to Halme et al.'s statement, we did not exaggerate the
143 interpretation of our results.

144

145 **Criteria of inclusion**

146 The use of p-values and other statistics to estimate an effect size is indeed possible although
147 relatively less often used in MA procedures than mean, SD and sample size. Several

148 reasons can explain why we did not use such data in our MA. First, the exact p, F or t-values
149 need to be available, which is not always the case (e.g. threshold values for probability).
150 Second, when those values are available, there could be two subsequent problems: (i) the
151 statistics could be extracted from a more or less elaborated model (i.e. with covariates), and
152 it is not advised to mix different sources of effect sizes in a meta-analysis (see Rosenberg et
153 al. 2000, p. 20); (ii) when several treatment classes are compared using a one-way analysis
154 of variance, the statistic simply tests if the means significantly differ from each other.
155 Therefore, it is impossible to transform the F or p-value of the ANOVA into an effect size,
156 because the effect size has to be computed from control and treatment means. Coming back
157 to the summary statistics is thus the only way to incorporate such results in a MA.
158 Consequently, contrary to Halme et al., we do not believe that we have overlooked "a great
159 deal of relevant literature" in our MA.
160 Another point raised by Halme et al. concerns the inclusion of the study by Sippola et al.
161 (2002): this paper compares old growth with 15 years-old stands, which were not considered
162 as "young regeneration phases" nor "clearfelling stands" in our protocol. We assume that our
163 selection protocol was restrictive enough regarding the number of studies finally included in
164 our MA; if we had been more restrictive in our inclusion criteria (i.e. excluding young stands),
165 we would have rejected this paper.

166

167 **Conclusions**

168 The paper we published does not aim at influencing European forest and conservation
169 policies in any way, but to provide decision-making tools based on scientific facts. Both
170 managed and unmanaged forests are needed to preserve European forest biodiversity, but
171 since there are many managed forests and very few old-growth ones, a special effort should
172 be allocated to create protected reserves, as suggested by Paillet et al. (2010).
173 Most of the comments of Halme et al. (except the suggestion to use p-values and other
174 statistics) would lead to reduce the number of comparisons, decrease the power of our meta-
175 analysis and weaken our conclusions. The methodological choices we made have intrinsic

176 limitations and cannot compensate for the weaknesses of the studies fed into it, but are
177 transparent: we chose a set of criteria to produce a standard protocol and followed it, as a
178 sound standard scientific practice. Then, we worked with the available data after following
179 our protocol. Moreover, we highlighted that future studies comparing biodiversity of managed
180 and unmanaged forests should better control for other sources of variation than management
181 and should systematically provide summary statistics. Many open questions remain and key
182 ideas for future research lay ahead.

183

184 **Literature cited**

185 Gurevitch, J., and L. V. Hedges. 2001. Meta-analysis. Combining the results of independent
186 experiments. Pages 347-369 in S. M. Scheiner, and J. Gurevitch, editors. Design and
187 Analysis of Ecological Experiments. Oxford University Press, New-York, USA.

188 Halme, P., T. Toivanen, M. Honkanen, J. S. Kotiaho, M. Monkkonen, and J. Timonen. this
189 issue. Biodiversity effects of forest management: flawed data scrutiny compromises
190 conclusions from meta-analysis. Conservation Biology.

191 Hurlbert, S. H. 1984. Pseudoreplication and the design of ecological field experiments.
192 Ecological Monographs **54**:187-211.

193 Hurlbert, S. H. 2009. The Ancient Black Art and Transdisciplinary Extent of
194 Pseudoreplication. Journal of Comparative Psychology **123**:434-443.

195 Koehnle, T. J., and J. C. Schank. 2009. An Ancient Black Art. Journal of Comparative
196 Psychology **123**:452-458.

197 Oksanen, L. 2001. Logic of experiments in ecology: is pseudoreplication a pseudoissue?
198 Oikos **94**:27-38.

199 Oksanen, L. 2004. The devil lies in details: reply to Stuart Hurlbert. Oikos **104**:598-605.

200 Paillet, Y., L. Bergès, J. Hjältén, P. Ódor, C. Avon, M. Bernhardt-Römermann, R.-J. Bijlsma,
201 L. De Bruyn, M. Fuhr, U. Grandin, R. Kanka, L. Lundin, S. Luque, T. Magura, S.
202 Matesanz, I. Mészáros, M.-T. Sebastià, W. Schmidt, T. Standovár, B. Tóthmérész, A.
203 Uotila, F. Valladares, K. Vellak, and R. Virtanen. 2010. Biodiversity differences

- 204 between managed and unmanaged forests: meta-analysis of species richness in
205 Europe. *Conservation Biology* **24**:101-112.
- 206 Rosenberg, M. S., D. C. Adams, and J. Gurevitch. 2000. *Metawin: Statistical Software for*
207 *Meta-Analysis*. Sinauer associates, Sunderland, Massachusetts.
- 208 Schank, J. C., and T. J. Koehnle. 2009. Pseudoreplication is a Pseudoproblem. *Journal of*
209 *Comparative Psychology* **123**:421-433.
- 210 Sippola, A. L., J. Siitonen, and P. Punttila. 2002. Beetle diversity in timberline forests: a
211 comparison between old-growth and regeneration areas in Finnish Lapland. *Annales*
212 *Zoologici Fennici* **39**:69-86.
- 213 Underwood, A. J. 1997. *Experiments in ecology: Their logical design and interpretation using*
214 *analysis of variance*. Cambridge University Press, Cambridge, United Kingdom.
- 215

216 Table 1: Distribution of individual studies used in Paillet et al. (2010) relative to Time Since
 217 Abandonment (TSA).

218

	TSA classes (years)				Total
	<50	50-75	75-100	>100	
All	21	32	22	14	89
Vascular plants ^a	10	4	8	1	23
Bryophytes	2	3	3	0	8
Lichens	0	3	5	2	10
Birds	3	0	4	0	7
Carabids	2	4	0	0	6
Saproxyllic beetles ^b	4	4	0	4	12
Non-saproxyllic beetles	0	4	0	2	6
Fungi	0	9	0	2	11

^a including ferns

^b including bark beetles

219