

0 Life-history traits and climatic responsiveness in noctuid moths 53

5 **Zsolt Végvári, Edit Juhász, János Pál Tóth, Zoltán Barta, Sándor Boldogh, Sándor Szabó and Zoltán Varga** 55

10 *Z. Végvári (vegvari@hnp.hu), Dept of Conservation Zoology, Hortobágy National Park Directorate, Univ. of Debrecen, HU-4024 Debrecen Sumen u.2, Hungary. – E. Juhász, S. Szabó and Z. Varga, Dept of Evolutionary Zoology and Human Biology, Univ. of Debrecen, HU-4032 Debrecen Egyetem tér 1., Hungary. – J. Pál Tóth, Res. Inst. for Viticulture and Oenology, Tarcal, Könyves Kálmán Street 54., Tarcal, HU-3915, Hungary. – Z. Barta, MTA-DE “Lendület” Behavioural Ecology Research Group, Dept of Evolutionary Zoology, Univ. of Debrecen, HU-4032 Debrecen Egyetem tér 1., Hungary. – S. Boldogh, Aggtelek National Park Directorate, HU-3758 Jósaváfi, Tengersizem oldal. 1., Hungary.* 60

15 Emergence phenology has been shown to advance considerably in the past decades in many lepidopterans. Noctuid moths (Noctuidae) constitute a species-rich family of lepidopterans with a large diversity of life history traits presumably driving climatic responsiveness. In our study we aim to assess the role of life-history and ecological traits in climatic responsiveness of noctuid moths, whilst controlling for phylogenetic dependence. We used a long-term dataset of European noctuid moths collected from a light-trap in northeastern Hungary. As the study site is located at the intersection of several biogeographical zones harbouring a large number of noctuid moth species, our dataset provides a unique possibility to investigate the moths' climatic sensitivity. To estimate the role of life-history traits and ecological factors in driving lepidopterans' response to climatic trends, we employed three proxies related to the species' ecology (habitat affinity, food plant specialization and food type) and two robust types of life-history traits (migration strategy and hibernation form). The degree of temporal shifts of various measures of emergence phenology was related to hibernation stage, food type and migration strategy. Large-scale phylogenetic relatedness exerted little constraint in all models fitted on each measure of phenology. Our results imply that noctuid moths overwintering as adults exhibited greater degrees of phenological shifts than species hibernating as larvae or pupae. It implies that moths hibernating as adults are forced to suspend activity in our climate and the prolongation of autumn activity might be the result of increased plasticity in flight periods. 70

30 During the past decades an increasing number of studies have reported on the effects of global climatic change on ecosystem structure and functioning (Parry et al. 2007, Walther 2010, Burrows et al. 2011). Climatic processes influence phenology, population dynamics and life-history traits in a broad range of organisms, including a large number of endangered taxa (Parmesan 2006, Forrest and Miller-Rushing 2010, Jenouvrier and Visser 2011). Specifically, evidence is accumulating that temporal shifts in insect phenology as a response to climatic alteration has accelerated over the past few decades (Singer and Parmesan 2010, Hodgson et al. 2011). For instance, emergence phenology has been shown to advance considerably in the past decades in many lepidopterans (Stefanescu et al. 2003, Hekkinen et al. 2010, Diamond et al. 2011). Besides, studies have recently detected that the annual number of generations of night-active moths has changed over the last decades (Altermatt 2010a, Kocsis and Hufnagel 2011, Gimesi et al. 2012). Global warming is assumed to be one of the most probable drivers of the increasing number of generations presumably by allowing more time for activity (Altermatt 2010a). Nocturnal moths have also been found to respond strongly to global warming by advancing spring flight periods but an immense number of species respond in a varied way to climatic changes in this taxa (Hodgson et al. 2011, Valtonen et al. 2011). 80

European noctuid moths constitute a representative taxonomic group for studying the effects of climatic trends on insect phenology, as they represent one of the most species-rich insect families subdivided into a large number of subfamilies and tribes characterised by a high diversity of life history, possibly related to the observed variation in climatic responsiveness (Karsholt and Razowski 1996, references in Supplementary material Appendix 1). A specific explanation for this large variation in climatic responsiveness may be that the ability of lepidopterans to track food plant phenology is assumed to be constrained by a number of ecological and life history factors (van Asch and Visser 2007, Altermatt 2010b, Breed et al. 2013). As a consequence, declining population sizes in butterflies (and also in moths) are often attributed to spatial mismatches between their distribution area and changing climatic envelopes, leading to regional or total extinctions of entire subspecies (Devictor et al. 2012). 90

The developmental stage in which individuals hibernate has already been shown to affect climatic responsiveness. For instance, species overwintering as adults were shown to exhibit higher degrees of phenological flexibility than 100

0 lepidopterans hibernating as larvae or pupae, as these taxa
 can emerge immediately after hibernation (Diamond et al.
 2011).

Recent studies have detected that climatic warming seems
 to enhance the range of migratory movements of some moth
 5 species, as migrant populations might track climatic trends
 faster owing to their large mobility (Sparks et al. 2007).

Evidence is accumulating that climatic responsiveness
 shows contrasting regional patterns across Europe, both in
 the frequency of supplementary generations and temporal
 10 patterns of emergence phenology of lepidopterans (Tobin
 et al. 2008, Altermatt 2010a, Pöyry et al. 2011, Kharouba
 et al. 2014).

In our study we aim to assess the role of life-history and
 ecological traits in climatic responsiveness in a species-rich
 15 lepidopteran family (Noctuidae) with a control for phylo-
 genetic dependence. To accomplish this objective, we used a
 long-term dataset of European noctuid moths collected by a
 light-trap in Jósvalfő (Aggtelek karst, northeastern Hungary,
 48.5°N, 20.51°E, 287 m a.s.l.), eastern Europe, where such
 20 investigations are still scarce.

Although a recent study has shown that many life-history
 traits might have important roles in climatic responsiveness
 in lepidopterans (Radchuk et al. 2013), we aim to assess only
 those which emerged as substantial predictors of climatic
 25 responsiveness in lepidopterans. Accordingly, we employed
 three proxies related to the species' ecology (habitat affini-
 ty, characterised as habitat specialist or generalist; food
 plant specialization and food type, defined as herbaceous or
 woody) and two types of life-history traits (migration strategy,
 30 and hibernation form) already shown to affect the climatic
 sensitivity of lepidopterans (Forrest and Miller-Rushing
 2010, Diamond et al. 2011).

Specifically, our hypotheses are as follows: 1) species with
 broader habitat range would be less responsive to climatic
 35 change, as these noctuids are less constrained by climate-
 induced changes in habitat phenology. 2) Food-generalist
 noctuid moths would exhibit smaller degrees of advance-
 ment in emergence phenology (as estimated by capture
 dates) than specialist species as they are less dependent on the
 40 phenology of one or a few specific food plants. 3) Migrant
 moths would track climatic variability to a greater degree
 than resident ones, as they have more developed sensitivity
 to weather conditions promoting flexible onsets of north-
 ward migration (Sparks et al. 2007, Alerstam et al. 2011).
 45 4) Finally, noctuids which hibernate as adults might track
 climatic trends faster than species hibernating as larvae or
 pupae as these moths can emerge immediately after finishing
 hibernation (Nyman et al. 2012).

50 Material and methods

Study area

55 The moth fauna of the Aggtelek karst area has been
 intensively surveyed over the last decades of the last century
 (Varga 1999, Szabó et al. 2007). We collected data dur-
 ing a long term light trap survey. The Jermy-type light trap
 was established in 1988 on the northeast-facing slope of
 60 the Tohonya Valley in Aggtelek National Park, northeast

Hungary (48.1°N, 20.51°E, 287 m). The trap equipped with
 a 125 W mercury bulb was activated in 1988 and operated
 in each year for a 20-year period until 2007 between March
 and November, spanning 4788 days in total. The study area
 provides a unique opportunity to investigate the moths'
 65 climatic sensitivity, as this part of the Carpathian basin is
 located in the intersection of several biogeographical zones
 harbouring 407 noctuid moth species, an unusually high
 number in central Europe (Varga 1999, 2012). Furthermore,
 the study area is surrounded by a broad variety of habitats
 70 hosting a set of noctuid taxa highly representative for the
 region. The light trap was located in an abandoned orchard,
 presently used as a semi-dry haymeadow, surrounded by scat-
 tered rests of karstic oak–hornbeam *Waldsteinio–Carpinetum*
 and xerothermic oak forests (*Corno–Quercetum pubescentis*)
 75 and hazel shrub *Coryletum avellanae* communities (Szabó
 et al. 2007).

Climate data

80 Weather data were obtained from the nearest meteorologi-
 cal station at Miskolc (48°1N, 20°4E), 60 km from the
 light-trap. Out of a numerous set of climatic variables, daily
 mean temperature and daily precipitation were available for
 the studied period. To assess temporal patterns in climatic
 85 predictors relevant for lepidopteran biology, we calculated
 monthly and seasonal (March to May for spring, June to
 August for summer, September to November for autumn
 and December to February for winter) means and variances
 of daily average temperature and monthly as well as seasonal
 90 precipitation sums and variances. Further, we computed
 temporal patterns in cumulative temperature, a biologically
 more relevant predictor of climatic responsiveness than tem-
 perature itself (Pöyry et al. 2011). The computation went as
 follows: we set a temperature limit of 5.0°C, and summed
 95 daily average temperature values above this limit for the
 following periods in each year: 1) 1 January to 31 March, 2)
 1 January to 31 April, 3) 1 January to 31 May, 4) 1 March
 to 31 May. To detect temporal trends in weather proxies,
 we fitted linear regressions on these variables as a function
 100 of years.

Noctuid data

105 In each year the light trap was operated from the first days of
 March until the first days of November (with a small, 1–2 day
 variation per year due to accidentally unfavourable weather
 conditions), totalling 4788 days (mean \pm SD = 239.4 \pm
 12 days per year). Captured specimens were collected on a
 110 daily basis and were stored in cotton layers or pinned until
 identification. Species-level identification was carried out by
 SSz and ZV and the data were entered into a database by
 recording date, taxon and number of individuals.

During the study period 128 045 individuals of 417
 noctuid taxa were collected. To acquire robust measures on
 115 climatic responsiveness, we used only the data of species with
 records from at least ten years spanning for at least 15 years,
 yielding 200 species in total. In the following step we retained
 only species with more than twenty individuals caught in
 total in every year, to be able to obtain reliable estimates of
 the annual population development. As a result, we used the
 121

0 records of 70 species in subsequent analyses (filtered species
 hereafter; Supplementary material Appendix 2). To test the
 robustness of the above filtering conditions, we removed
 noctuids with less than ten distinct capture dates per year
 from the set of 200 species and repeated all analyses.

5 To test the effects of first and final operation dates of
 the light trap, we calculated the distribution of capture dates
 of filtered species. This shows that the percentage of first and
 last capture dates within seven days from first and last opera-
 tion dates, respectively, were less than 0.1%. Therefore it is
 10 unlikely that between-year variation in first and final opera-
 tion dates can influence phenological patterns.

To derive standard measures of climatic responsiveness for
 each species, we calculated slopes of linear regressions fitted
 on years using the following estimates of emergence phenol-
 15 ogy as response variables: 1) first, 2) median and 3) last julian
 dates (day-of-year number) of capture and 4) activity length
 (defined as the number of days between first and last capture
 dates), calculated using corresponding capture dates. We
 refer to these response variables as climatic responsiveness
 20 of first, median and last julian days as well as activity length.
 High slopes from these regressions indicate that a given
 species can respond faster to climatic changes. The linear
 regression slopes were then applied in phylogenetic least
 squares models as response variables to investigate which
 25 ecological and life history traits influence climatic respon-
 siveness. To test the robustness of the emergence predictors,
 we applied the approach proposed by van Strien et al. (2008)
 to our dataset. Accordingly, we calculated 1) the date of the
 25th individual captured (FA25) and 2) the date of the first
 30 25% of all observed individuals in the entire flight period
 (FP25) for each species in the filtered dataset. In the next
 step we calculated Pearson's correlation coefficients medium
 capture dates and only FP25 which revealed significant rela-
 tionships ($r = 0.73$, $p < 0.03$) between medium capture
 35 dates and those proposed by van Strien et al. (2008).

Ecological and life-history predictors

We obtained data on ecological variables and life-history
 40 traits from the series of Noctuidae Europae (Ronkay
 et al. 2001, Hacker et al. 2002, Fibiger et al. 2009, 2010).
 Habitat affinity was classified as specialist or generalist.
 Feeding strategy was characterized by diet specialization
 (categorised as food specialist or food generalist) and food
 45 type (classified according to the food type of caterpillars:
 herbaceous or woody). Migration strategy was categorised
 as resident and migratory. Hibernation type was classified as
 imago, pupal and larval stages (Varga et al. 2005).

All used data on the species' life history and ecology are
 50 presented in the Supplementary material Appendix 2.

Phylogenetic and statistical analyses

In the first step we constructed a composite tree (Supplementary
 55 material Appendix 3) using the data of the most recent molecu-
 lar survey applying gradual branch lengths (Mitchell et al. 2006,
 Zahiri et al. 2011). To account for the lack of knowledge of
 species-level tree topology, we included only subfamily-level
 information provided by the most recent molecular phyloge-
 60 netic analyses (Mitchell et al. 2006, Zahiri et al. 2011).

To evaluate the degree of association between climatic
 61 responsiveness measures and ecological and life-history
 predictors, we fitted phylogenetic generalized least squares
 (PGLS) models controlling for the non-independence
 65 among taxa by incorporating a variance-covariance matrix
 that represents their phylogenetic relationships (Martins
 and Hansen 1999, Pagel 1997, 1999). The degree of
 phylogenetic dependence (λ) was set to the most appropriate
 value obtained for each model by likelihood ratio statistics
 (Freckleton et al. 2002). In the following step we calcu-
 70 lated the strength of phylogenetic relatedness in each trait
 separately by quantifying the maximum-likelihood value
 of λ with its statistical significance estimated by likelihood
 ratio tests (i.e. by comparing the log-likelihood of the model
 in which λ is set to the most appropriate value to the log-
 75 likelihood of a model in which $\lambda = 0$). As recent studies
 on the phylogeny of Noctuidae have suggested a number
 of major changes in the structure of the phylogenetic tree
 (references provided in the Supplementary material Appen-
 80 dix 4), we repeated assessing phylogenetic signal in response
 and predictor variables by formulating linear mixed mod-
 els that include only genus nested in subfamily as a random
 term, and compared model performance by ANOVA tests.

After model fitting, the relative importance of life-history
 85 traits and ecological variables was calculated using informa-
 tion-theoretic model-comparison (Burnham and Anderson
 2002). First we obtained the values of Akaike's informa-
 tion criterion corrected for small sample size (AIC_c) and the
 corresponding Akaike weight of each model (ω). Second,
 we selected models with substantial support ($\Delta_i = AIC_i$
 90 $- AIC_{\min} < 2.0$) and calculated model-averaged param-
 eter estimates (β) and unconditional standard errors (SE_{β} ;
 Burnham and Anderson 2002) of each variable by the sums
 of their Akaike weights across all models with substantial
 support containing the given predictor. Akaike differences in
 95 the range 0–2 indicate substantial level of empirical support
 of a given model, whereas $\Delta_i > 7$ implies very weak support
 (Burnham and Anderson 2002).

For all analyses we used the R statistical computing
 100 environment including its 'ape' (Paradis et al. 2011), 'geiger'
 (Harmon et al. 2009), 'MuMIn' (Barton 2011) and 'nlme'
 (Pinheiro et al. 2009) packages.

Results

Climatic patterns

105 During the 20 years of data collection monthly mean temper-
 ature has significantly increased only in June ($b = 0.140^\circ\text{C}$
 110 year^{-1} , $F_{1,17} = 7.292$, $p = 0.015$, Fig. 1). Seasonal mean and
 cumulative temperature has not shown significant trends.
 Further, we detected no clear temporal patterns in variance
 of monthly temperature and all precipitation metrics (results
 115 not shown).

Temporal trends in capture data

Noctuids showed a considerable amount of variation in
 slopes of various estimates of emergence phenology as a
 function of years. Out of the 70 taxa in our study, 38 shifted
 121

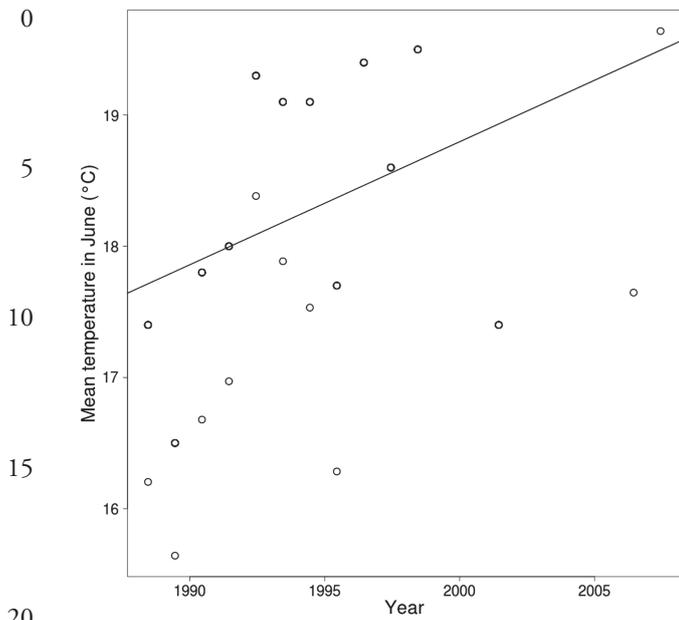


Figure 1. June mean temperature as a function of years during the study.

their first capture to earlier dates with four of them exhibiting significant advance. Median capture dates, indicative of the climatic responsiveness of the bulk of the population, showed similar patterns with 34 noctuids emerging earlier but the trend was only significant for a single species. Last julian dates of annual records have advanced in 27 taxa, four of which were significant (Supplementary material Appendix 2, Fig. 2).

In contrast, first capture of 32 noctuids shifted to later dates with only a single species showing significant patterns. Median capture date was detected to change to later dates in 36 noctuids including three species with significant trends. Last capture dates of 43 noctuids have shifted later, five of

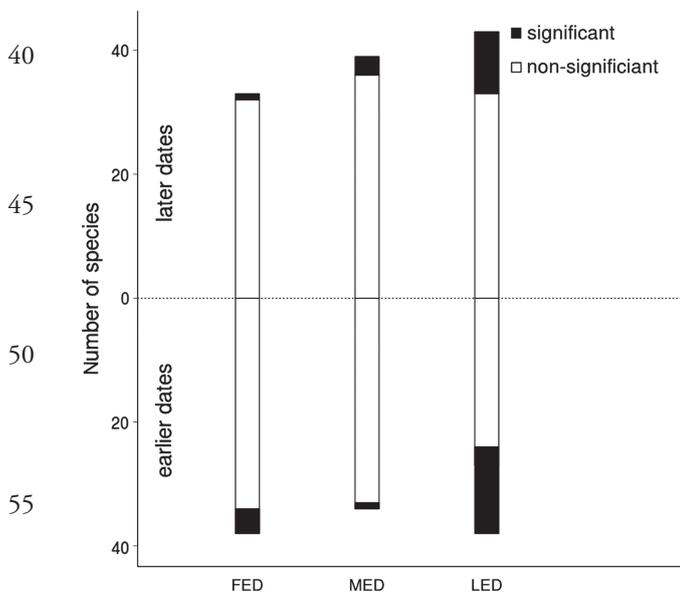


Figure 2. Number of species shifting emergence phenology to earlier and later dates.

Table 1. λ -statistics for phylogenetic signal in the investigated response variables and predictors. $\lambda = 0$ indicates no phylogenetic signal.

Variable	λ	Likelihood ratio	p
Slope of first capture date	0.000	0.000	0.999
Slope of median capture date	0.000	0.000	0.999
Slope of last capture date	0.000	0.000	0.999
Slope of activity length	0.000	0.000	0.997
Migration type	0.000	0.000	0.999
Hibernation type	0.000	0.000	0.422
Food specialisation	0.543	0.898	0.343
Food type	0.811	2.249	0.134
Habitat affinity	0	0.001	0.98

which exhibited significant temporal patterns. Although activity length shortened in 28 species with significant trends in three species, it became longer in 42 noctuids including significant relationships for nine taxa (Supplementary material Appendix 2).

On average, noctuids shifted their first capture dates by 4.705 (SE = 0.902) days to later dates after 1998 than before this year across all species considered. Similarly, median capture occurred 4.525 (SE = 2.210) days later in the second period than in the first one. In contrast, overall last emergence shifted by 7.547 (SE = 1.325) days to later dates, representing a delay almost twice as long as the delay at the beginning of the flight season. As a result, activity length increased by 2.842 (SE = 1.151) days calculated over the filtered set of species.

Ecological and life history correlates of phenological change

Phylogenetic signal (λ) was zero for all measures of all response and proxy variables (Table 1). Quantitatively similar results were obtained by linear mixed models which showed no significant effect of genus nested within subfamily ($p = 0.067$).

Responsiveness of first capture dates was significantly affected by food type and migration strategy – parameter estimates of which had confident intervals not overlapping with zero. Noctuids with larvae feeding primarily on woods are shifting first emergence to later dates while species feeding on herbaceous plants advance first capture dates ($\Sigma = 0.826$; Table 2, 3; Supplementary material Appendix 5).

Our results show that migratory strategy has emerged as the second most important proxy of climatic responsiveness measured as first capture dates based on its Akaike sums in PGLS-models fitted on first capture dates ($\Sigma = 0.749$; Table 2, 3; Supplementary material Appendix 5): resident noctuids advanced climatic responsiveness measured as first capture dates to a greater degree than migratory species. Parameter estimates for all other variables have confidence intervals overlapping with zero, therefore we consider their effects as non-significant.

Considering last emergence phenology hibernation stage has emerged with the highest Akaike scores calculated over supported models fitted on the slope of last capture dates ($\Sigma = 1.0$, Table 2). Hibernation stage was the only predictor

0 Table 2. Akaike sums, model-averaged parameter estimates (β), standard errors (SE_u) and 95% confidence intervals for predictors across best supported models ($\Delta_i < 2.0$). 61

	Slope of first emergence date	Slope of median emergence date	Slope of last emergence date	Slope of activity length	
5	Food type $\Sigma = 0.826$ $\beta(SE_u) = -0.395 (0.145)$ 95% CI = (-0.678; -0.111)	Migration strategy $\Sigma = 0.483$ $\beta(SE_u) = -0.497 (0.575)$ 95% CI = (-1.625; 0.631)	Hibernation stage $\Sigma = 1.0$ $\beta(SE_u) = -2.674 (0.543)$ 95% CI = (-3.737; -1.61)	Hibernation stage $\Sigma = 1.0$ $\beta(SE_u) = -3.018 (0.608)$ 95% CI = (-4.21; -1.826)	65
10	Migration strategy $\Sigma = 0.749$ $\beta(SE_u) = -0.547 (0.243)$ 95% CI = (-1.023; -0.07)	Food type $\Sigma = 0.404$ $\beta(SE_u) = -0.376 (0.365)$ 95% CI = (-1.091; 0.339)	Migration strategy $\Sigma = 0.246$ $\beta(SE_u) = -0.227 (0.406)$ 95% CI = (-1.023; 0.57)	Migration strategy $\Sigma = 0.282$ $\beta(SE_u) = 0.265 (0.467)$ 95% CI = (-0.651; 1.18)	70
15	Habitat affinity $\Sigma = 0.296$ $\beta(SE_u) = 0.252 (0.15)$ 95% CI = (-0.099; 0.548)	Habitat affinity $\Sigma = 0.299$ $\beta(SE_u) = 0.206 (0.36)$ 95% CI = (-0.499; 0.912)	Food type $\Sigma = 0.209$ $\beta(SE_u) = -0.217 (0.301)$ 95% CI = (-0.807; 0.373)	Habitat affinity $\Sigma = 0.213$ $\beta(SE_u) = -0.212 (0.287)$ 95% CI = (-0.774; 0.35)	75
20	Food specialisation $\Sigma = 0.212$ $\beta(SE_u) = 0.224 (0.165)$ 95% CI = (-0.099; 0.548)	Food specialisation $\Sigma = 0.293$ $\beta(SE_u) = 0.052 (0.404)$ 95% CI = (-0.74; 0.843)	Food specialisation $\Sigma = 0.183$ $\beta(SE_u) = 0.151 (0.293)$ 95% CI = (-0.424; 0.725)	Food type $\Sigma = 0.195$ $\beta(SE_u) = 0.099 (0.339)$ 95% CI = (-0.564; 0.763)	80
20	Hibernation stage $\Sigma = 0.151$ $\beta(SE_u) = 0.386 (0.303)$ 95% CI = (-0.208; 0.981)	Hibernation stage $\Sigma = 0.278$ $\beta(SE_u) = 0.097 (0.755)$ 95% CI = (-1.384; 1.577)	Habitat affinity $\Sigma = 0.142$ $\beta(SE_u) = -0.018 (0.252)$ 95% CI = (-0.512; 0.475)	Food specialisation $\Sigma = 0.186$ $\beta(SE_u) = -0.06 (0.331)$ 95% CI = (-0.708; 0.588)	80

of last capture dates and activity length in the best supported models: species hibernating as imagos shifted last capture to later dates and prolonged activity length to a greater degree than noctuids hibernating as larvae or pupae (Table 2, Fig. 3, 4; Supplementary material Appendix 7).

We obtained qualitatively similar results for species with more than ten capture dates per year (Supplementary material Appendix 6).

Discussion

In this study we aimed to 1) describe the extent of phenological responses presumably caused by ongoing climate change in a species-rich group of noctuid moths and 2) investigate the possible ecological and life history correlates of these responses. We believe that our results are based on a unique combination of methodological elements. First, our dataset was collected over 20 years applying highly standardised collection methodology. Second, the single collection site allows high-level replication of species-specific climatic responsiveness of large number of noctuid taxa

allowing multispecies analyses of 70 species after rigorous filtering. Third, we calculated multiple phenological estimates in order to avoid collection bias primarily affecting first emergence phenology.

Considering noctuid phenology, we detected no clear patterns in changes of emergence phenology, which might be related to the diversity in a number of life history traits of the studied species, as shown by recent investigations (Altermatt 2010a or b????). Additionally, complex trends of emergence phenology can also be a result of our finding that significant trends in climatic patterns were only found for June temperature, which is in line with a number of recent studies showing no clear temperature trends after 1990 (reviewed by Stocker et al. 2013). Although Hungary has experienced considerable increase of annual mean temperature over the past century (Molnár V. et al. 2012), our study covered only two decades and was located in an area where the foothills of the Carpathian Mountains might have had a buffering effect.

Our findings show that the degree of temporal shifts of various measures of emerging phenology was primarily related to hibernation stage, food type and migration strategy.

Table 3. Supported models ($\Delta_i < 2.0$) fitted on all types of response variables.

Response	Model formula	DF	AICc	Δ_i	w	
Slope of first emergence date	food type + migration strategy	4	123.811	0	0.279	110
	food type + habitat affinity + migration strategy	5	124.904	1.092	0.161	
	food type	3	125.321	1.509	0.131	
Slope of median emergence date	migration strategy	3	246.815	0	0.123	115
	food type	3	247.018	0.203	0.111	
	food type + migration strategy	4	247.695	0.88	0.079	
	hibernation stage	4	247.858	1.043	0.073	
	food specialisation	3	247.881	1.067	0.072	
Slope of last emergence date	habitat affinity	3	247.996	1.181	0.068	115
	habitat affinity + migration strategy	4	248.586	1.771	0.051	
	hibernation stage	4	203.804	0	0.395	
Slope of activity length	hibernation stage	4	219.033	0	0.362	121
	hibernation stage + migration strategy	5	220.947	1.914	0.139	

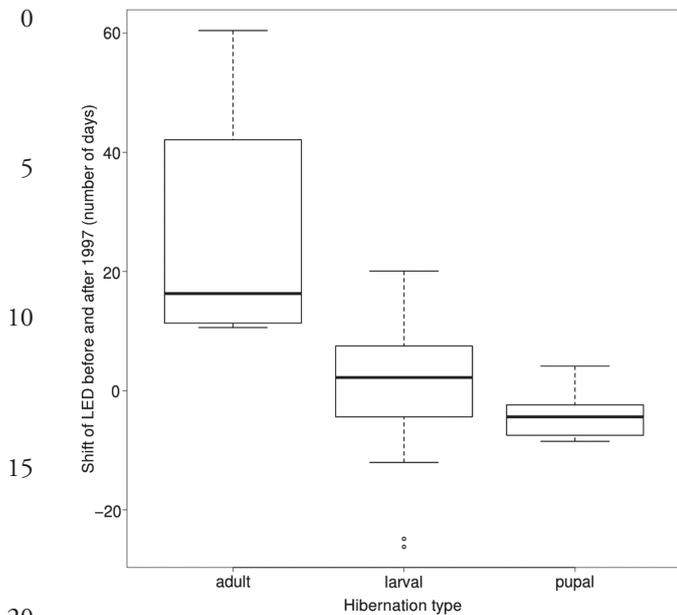


Figure 3. Shift of last capture date as a function of hibernation stage in noctuid moths.

Specifically, hibernation stage is the strongest predictor of climatic responsiveness measured as last capture dates and activity length in noctuid moths, similarly to the results of Diamond et al. (2011). This might indicate that noctuid moths overwintering as adults exhibited greater degrees of phenological shifts than species hibernating as larvae or pupae. This finding extends the results of a recent study reporting that species with more advanced overwintering stages have experienced relatively greater advances in their date of first appearance (Diamond et al. 2011). However, the difference between the reactions of the bulk of the population and last active individuals to climatic trends seems to be

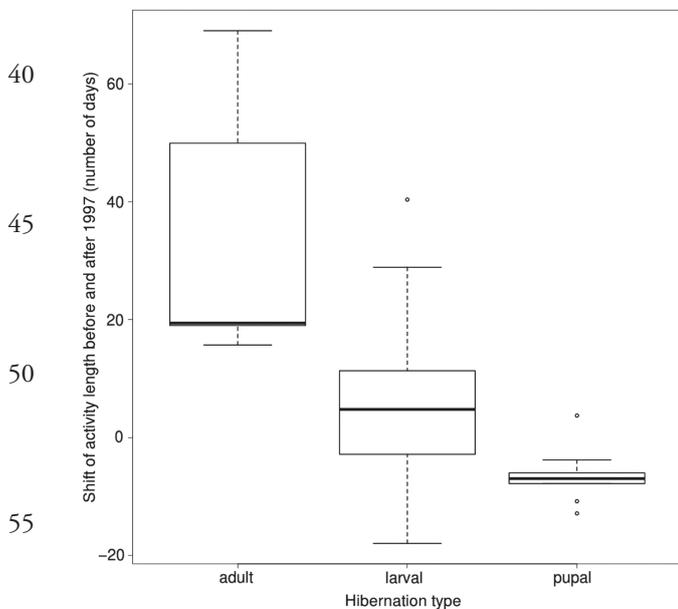


Figure 4. Shift of activity length as a function of hibernation stage in noctuid moths.

different in our case. Since noctuids hibernating as adults in our database showed no shifts in median emergence to earlier dates, the last individuals of these moths have prolonged capture dates over the study period in comparison to species hibernating in larval or pupal form. Prolonged autumn activity may allow longer time for replenishing fat reserves before hibernation which can increase survival chances and improve reproductive success. Moreover, all of these species are winter-active moths in the southern part of their distribution areas implying that noctuid taxa hibernating as imago are forced to suspend activity in our climate and the prolonged autumn activities of the last emerging individuals might be the result of enhanced temporal plasticity of the timing of autumn emergence (László et al. 2000).

Food type was included in best supported models for climatic responsiveness measured as first capture dates, similarly to previous investigations: noctuids feeding on herbaceous plants advanced first capture dates to a greater degree than species connected to woody materials (Diamond et al. 2011, Altermatt 2010a or b????). Although a previous study showed that herbaceous plants were more likely than woody plants to advance phenology (Walther et al. 2002), our results imply that moths feeding on herbaceous plants are able to track plant phenology faster than noctuids feeding on woody plants which evokes the following hypotheses. First, as herbaceous plants are shorter-lived on average than woody taxa, their herbivores are forced to synchronise phenology to a greater degree than moth larvae feeding on woody plants. Second, woody plants are on average larger than herbaceous species implying higher levels of phenological variance of plant parts consumed by herbivores in individual trees which allows for a wider range of phenological synchronisation. However, the complexity of climatic responses driven by trophic interactions is enhanced by the fact that the highly diverse family of Noctuidae is linked to a great variety of larval food plants (references in Supplementary material Appendix 1). For instance, the majority of these species is connected to forested habitats and majority of spring- and autumn-active moths are linked to woody food plants (details shown in Supplementary material Appendix 2; references in Supplementary material Appendix 1). A large group of genera (Supplementary material Appendix 2) use a wide variety of food plants which are most often herbaceous (grasses for mostly monovoltine *Apamea*, and often Urticaceae for the phenologically more plastic Plusiinae), or partly herbaceous and woody (Acronictinae) (Supplementary material Appendix 2).

Migration strategy has emerged as a supported predictor of climatic responsiveness. Migrant noctuids have shifted first captures to later dates than resident moths which implies that migratory species arriving from southern regions may detect warming spring weather later than locally hibernating taxa (Alerstam et al. 2011). Indeed, migrant moths need time to reach central Europe and thus are outcompeted in advancements of climatic responsiveness measured as first capture dates by locally wintering moths hibernating as imagoes.

Throughout our analyses we detected no phylogenetic relatedness on the subfamily level in climatic responsiveness of noctuid moths. As the existing polytomies are largely unresolved, this implies that the absence of phylogenetic signal in climatic responsiveness of noctuid moths

- 0 can be considered as the result of poor resolution of clades
rather than missing phylogenetic relationships. Since the
majority of noctuid taxa in our dataset belongs to the tribe
Hadenini previously merged into the subfamily Noctui-
nae by Lafontaine and Schimdt (2011), for which recent
5 investigations on evolutionary relatedness do not provide
better phylogenetic resolutions, we were not able to detect
phylogenetic signal on finer scales (Zahiri et al. 2013).
This warrants further research on phylogenetic related-
ness among noctuid moths to be able to detect if climatic
10 responsiveness is independent of phylogenetic relatedness
in noctuid species as shown by a number of investigations
showing that phenological shifts are not driven by com-
mon descent in a broad range of taxa (Végyári et al. 2010,
Diamond et al. 2011). However, the lack of phylogenetic
15 signal in emergence phenology of the species in our data-
set can also be an outcome of phenotypic plasticity which
drives phenological patterns as a response to climatic vari-
ability, suggested by the findings of recent studies across a
number of taxonomic levels (Hoffmann and Sgrò 2011).
[AQ4] 20 The similar results of Diamond et al. (2011) showing that
phylogenetic relatedness among species predicted very little
variation in butterflies' phenological responses imply that
climatic responsiveness in lepidopterans can be treated as a
species-specific reaction rather than driven by evolutionary
25 relatedness (Altermatt 2010a).
Our study implies recommendations for lepidopteran
conservation, as a reflection to current declining trends
in a number of insect species on a global scale. We found
that noctuid moths hibernating as adults are able to track
30 climatic trends faster than species overwintering in less
developed stages which suggest higher declining probabili-
ties in these taxa, considering current climatic predictions
(Stocker et al. 2013). Our result that resident moths are
responding to ongoing climatic patterns to a greater degree
35 than migratory taxa, calls for more intensive surveys of
migrating lepidopteran populations.
In summary, we have demonstrated in a diverse set of
noctuid moth species that temporal trends in emergence
phenology (which is associated with a measure of tempera-
40 ture increase in the study area) is primarily related to hiber-
nation stage, diet specialisation and migration strategy.
- Acknowledgements* – This research was supported by the European
Union and the State of Hungary, co-financed by the European
45 Social Fund in the framework of TÁMOP-4.2.4.A/2–11/1–2012-
0001 'National Excellence Program'. The support of the nature con-
servation authorities of Hungary is highly appreciated. ZB was
supported by a Hungarian Research Fund grant (OTKA K75696).
- 50 **References**
- Alerstam, T. et al. 2011. Convergent patterns of long-distance
nocturnal migration in noctuid moths and passerine birds.
– Proc. R. Soc. B 278: 3074–3080.
55 Altermatt, F. 2010a. Climatic warming increases voltinism in
European butterflies and moths. – Proc. R. Soc. B 277:
1281–1287.
Altermatt, F. 2010b. Tell me what you eat and I'll tell you when
you fly: diet can predict phenological changes in response to
60 climate change. – Ecol. Lett. 13: 1475–1484.
- Barton, K. 2011. MuMIn: Multi-model inference. – R package ver. 1.0.0. 61
Breid, G. A. et al. 2013. Climate-driven changes in northeastern US
butterfly communities. – Nature Climate Change 3: 142–145.
Burnham, K. P. and Anderson, D. R. 2002. Model selection and
65 multi-model inference: a practical information-theoretic
approach. – Springer.
Burrows, M. T. et al. 2011. The pace of shifting climate in marine
and terrestrial ecosystems. – Science 334: 652–655.
Devictor, V. et al. 2012. Differences in the climatic debts of birds
and butterflies at a continental scale. – Nature Climate Change 70
2: 121–124.
Diamond, S. E. et al. 2011. Species' traits predict phenological
responses to climate change in butterflies. – Ecology 92:
1005–1012.
Fibiger, M. et al. 2009. Noctuidae Europaeae, Pantheinae–
75 Bryophilinae. – Entomological Press, Sorø.
Fibiger, M. et al. 2010. Noctuidae Europaeae, Rivulinae–
Phytometrinae, and Micronoctuidae, including Supplement to
Noctuidae Europaeae. – Entomological Press Sorø.
Forrest, J. and Miller-Rushing, A. J. 2010. Toward a synthetic
understanding of the role of phenology in ecology and evolu-
80 tion. – Phil. Trans. R. Soc. B 365: 3101–3112.
Freckleton, R. P. et al. 2002. Phylogenetic analysis and comparative
data: a test and review of evidence. – Am. Nat. 160: 712–726.
Gimesi, L. et al. 2012. The effect of climate change on the phenol-
ogy of moth abundance and diversity. – Appl. Ecol. Environ.
85 Res. 10: 349–363.
Hacker, H. et al. 2002. Noctuidae Europaeae, Hadeninae I.
– Entomological Press, Sorø, pp. 419.
Harmon, L. et al. 2009. geiger: analysis of evolutionary diversifica-
tion. – R package ver. 1.3-1.
Heikkinen, R. K. et al. 2010. Assessing the vulnerability of [AQ3]
European butterflies to climate change using multiple criteria. 90
– Biodivers. Conserv. 19: 695–723.
Hodgson, J. A. et al. 2011. Predicting insect phenology across space
and time. – Global Change Biol. 17: 1289–1300.
Hoffmann, A. A. and Sgrò, C. M. 2011. Climate change and [AQ4]
evolutionary adaptation. – Nature 470: 479–485. 95
Jenouvrier, S. and Visser, M. E. 2011. Climate change, pheno-
logical shifts, eco-evolutionary responses and population viabil-
ity: toward a unifying predictive approach. – Int. J.
Biometeorol. 5: 905–919.
Karsholt, O. and Razowski, J. 1996. The Lepidoptera of Europe:
a distributional checklist. – Apollo Books. 100
Kharouba, H. M. et al. 2014. Predicting the sensitivity of
butterfly phenology to temperature over the past century.
– Global Change Biol. 20: 504–514.
Lafontaine, J. D. and Schmidt, B. C. 2011. Additions and correc-
105 tions to the check list of the Noctuoidea (Insecta, Lepidoptera)
of North America north of Mexico. – ZooKeys 149: 145.
Kocsis, M. and Hufnagel, L. 2011. Impacts of climate change on
Lepidoptera species and communities. – Appl. Ecol. Environ.
Res. 9: 43–72.
László, G. M. et al. 2000. On the genesis of the Himalayan-Sino
Pacific Thyatiridae (Lepidoptera) fauna, with special reference
110 to Taiwan. – In: Biodiversity across the Taiwan Strait. Nat
Mus. Nat. Sci., Taichung, poster volume, 65.
Martins, E. P. and Hansen, T. F. 1999. Phylogenies and the
comparative method: a general approach to incorporating
phylogenetic information into the analysis of interspecific data
– Am. Nat. 153: 448–448. 115
Mitchell, A. et al. 2006. Systematics and evolution of the cutworm
moths (Lepidoptera : Noctuidae): evidence from two protein-
coding nuclear genes. – Syst. Entomol. 31: 21–46.
Molnár, A. et al. 2012. Pollination mode predicts phenological
response to climate change in terrestrial orchids: a case study
120 from central Europe. – J. Ecol. 100: 1141–1152.

0	Nyman, T. et al. 2012. Climate-driven diversity dynamics in plants and plant-feeding insects. – <i>Ecol. Lett.</i> 15: 889–898.	
	Pagel, M. 1997. Inferring evolutionary processes from phylogenies. – <i>Zool. Scripta</i> 26: 331–348.	
	Pagel, M. 1999. Inferring the historical patterns of biological evolution. – <i>Nature</i> 401: 877–884.	
5	Paradis, E. et al. 2011. Package ‘ape’: analysis of phylogenetics and evolution. Publisher????	
	Parmesan, C. 2006. Ecological and evolutionary responses to recent climate change. – <i>Annu. Rev. Ecol. Evol. Syst.</i> 37: 637–669.	
	Parry, M. L. et al. 2007. Climate change 2007: impacts, adaptation and vulnerability. Intergovernmental Panel on Climate Change. – Cambridge Univ. Press.	
10	Pinheiro, J. et al. 2009. Nlme package. Publisher????	
	Pöyry, J. et al. 2011. Climate-induced increase of moth multivoltinism in boreal regions. – <i>Global Ecol. Biogeogr.</i> 20: 289–298.	
15	Radchuk, V. et al. 2013. Each life stage matters: the importance of assessing the response to climate change over the complete life cycle in butterflies. – <i>J. Anim. Ecol.</i> 82: 275–285.	
	Ronkay, L. et al. 2001. Noctuidae Europaeae. Hadeninae II. – Entomological Press, Sorø, pp. 452.	
20	Singer, M.C. and Parmesan, C. 2010. Phenological asynchrony between herbivorous insects and their hosts: signal of climate change or pre-existing adaptive strategy? – <i>Phil. Trans. R. Soc. B</i> 365: 3161–3176.	
	Sparks, T. H. et al. 2007. Increased migration of Lepidoptera linked to climate change. – <i>Eur. J. Entomol.</i> 104: 139–143.	
25	Stefanescu, C. et al. 2003. Effects of climatic change on the phenology of butterflies in the northwest Mediterranean Basin. – <i>Global Change Biol.</i> 9: 1494–1506.	
	Stocker, T. F. et al. 2013. Climate change 2013: the physical science basis. Working Grp I Contrib. to the 5th Assess. Rep. of the Intergovernmental Panel on Climate Change. Summary for Policymakers (IPCC, 2013). Publisher???	
30		
	Szabó, S. et al. 2007. Long-term light trap study on the macro-moth (Lepidoptera : Macroheterocera) fauna of the Aggtelek National Park. – <i>Acta Zool. Acad. Sci. Hung.</i> 53: 257–269.	61
	Tobin, P. C. et al. 2008. Historical and projected interactions between climate change and insect voltinism in a multivoltine species. – <i>Global Change Biol.</i> 14: 951–957.	65
	Valtonen, A. et al. 2011. Environmental controls on the phenology of moths: predicting plasticity and constraint under climate change. – <i>Oecologia</i> 165: 237–248.	
	van Asch, M. and Visser, M. E. 2007. Phenology of forest caterpillars and their host trees: the importance of synchrony. – <i>Annu. Rev. Entomol.</i> 52: 37–55.	70
	van Strien, A. J. et al. 2008. Bias in phenology assessments based on first appearance data of butterflies. – <i>Oecologia</i> 156: 227–235.	
	Varga, Z. 1999. The Lepidoptera of the Aggtelek National Park. – In: Mahunka S. (ed.), The fauna of the Aggtelek National Park II. Hung. Nat. Hist. Mus., Budapest, pp. 443–504.	75
	Varga, Z. 2012. Magyarország nagylepkéi – Macrolepidoptera of Hungary. – Heterocera Press, Budapest.	
	Varga, Z. et al. 2005. Checklist of the fauna of Hungary. Macrolepidoptera. – Hung. Nat. Hist. Mus., Budapest	80
	Végvári, Z. et al. 2010. Life history predicts advancement of avian spring migration in response to climate change. – <i>Global Change Biol.</i> 16: 1–11.	
	Walther, G. R. 2010. Community and ecosystem responses to recent climate change. – <i>Phil. Trans. R. Soc. B</i> 365: 2019–2024.	85
	Walther et al. 2002	[AQ1]
	Zahiri, R. et al. 2011. A new molecular phylogeny offers hope for a stable family level classification of the Noctuoidea (Lepidoptera). – <i>Zool. Scripta</i> 40: 158–173.	
	Zahiri, R. et al. 2013. Relationships among the basal lineages of Noctuidae (Lepidoptera, Noctuoidea) based on eight gene regions – <i>Zool. Scripta</i> 42: 488–507.	90
	Supplementary material (available online as Appendix oik.01655 at < www.oikosjournal.org/readers/appendix >).	95
35	Appendix 1–7.	
40		100
45		105
50		110
55		115
60		121