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In stable, unmanaged grasslands local factors are more important than landscape-level factors in shaping spider assemblages

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A B S T R A C T

Previous studies reported that landscape-level factors are vital to support diversity of spiders in strongly modified arable lands and disturbed habitats such as managed semi-natural grasslands. Cropland management (ploughing, fertilization, and pest management) and agricultural practices (mowing and grazing) destroy and/or modify regularly the spider assemblages; thus, continuous recolonization from the surrounding landscape is vital to sustain the species pool. On the contrary, we hypothesized that in unmanaged grasslands, the spider assemblages are stable and the importance of recolonization is limited, the local factors become much more important drivers in shaping spider assemblages than landscape-level factors. We tested the importance of local and landscape-level factors on the abundance and species richness of spiders in unmanaged grasslands. At the landscape-level, we found that only the isolation had significant effect on the total abundance, on the abundance of hunting and habitat specialist species, and on the abundance of a frequent species (Gnaphosa mongolica). At the local scale, however, four out of five studied factors influenced significantly the species richness and abundance of spider assemblages and the abundance of two frequent species (Alopecosa psammophila, Berlandia cineria). Species richness and abundance increased by plant cover, litter cover, and patch size, while decreased by bare ground cover. We found that in unmanaged grasslands, the local factors had vital role in maintaining the spider species richness; this is just the opposite conclusion that was earlier reported for agricultural ecosystems, where landscape-level effects had crucial role providing the species for continuous recolonization.

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1. Introduction

Prior to agricultural intensification, natural and semi-natural grasslands were one of the most diverse habitat types in Europe (Batáry et al., 2008). Since the second half of the last century, the increasing agricultural activity has been an important driver of biodiversity loss in these grasslands (Tscharntke et al., 2005; Krauss et al., 2010; Hooftman and Bullock, 2012; Dengler et al., 2014). Low-intensity grassland systems in Central and Eastern Europe maintain a diverse and unique flora and fauna (Varga, 1997; Török et al., 2000). During the last decades due to the increase of large-scale farming and abandonment of the traditional management practices, the Central and Eastern European grasslands also become highly fragmented and endangered (Horváth et al., 2009; Buchholz, 2010).

Grasslands play an important role in the maintenance of biodiversity in cultivated landscapes by providing habitats and/or refuges for many species (Jannneret et al., 2003; Woodcock et al., 2005; Horváth et al., 2013). Survival of the majority of grassland arthropod species in cultivated landscapes primarily depends on the quality of habitats, but also depends on the surrounding landscape (Jannneret et al., 2003). To understand the relationship between agricultural activity and grassland biodiversity, it is essential to investigate the effects at different spatial scales (e.g., the local scale and the landscape-level scale) in these habitats (Tscharntke et al., 2005, 2012; Batáry et al., 2008).

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Spiders are one of the most abundant and species rich generalist predators. They contribute to the biodiversity in natural and agricultural ecosystems (Wise, 1993). The occurrence of spiders mostly depends on local factors such as vegetation structure, vegetation composition, microclimate conditions, and prey availability (Heikkinen and MacMahon, 2004; Horváth et al., 2005; Batáry et al., 2008; Schirmer et al., 2011). Landscape-level factors such as percentage of grasslands, percentage of non-crop habitats, and landscape diversity also influence spider assemblages (Clough et al., 2005; Horváth et al., 2013; Schmidt et al., 2008).

Moreover, the human disturbance (urbanization and fragmentation), natural disturbance (e.g., fire, flood, and drought), and management regime (mowing, grazing, and burning) are also important factors (Cattin et al., 2003; Horváth et al., 2009, 2012; Malumbres-Olarte et al., 2014). To understand the changes in structure and composition of spider assemblages in natural and semi-natural habitats, it is important to investigate the effects of environmental factors at both local and landscape-level scales (Batáry et al., 2008).

In the previous studies, the effects of local and landscape-level factors on spiders were mainly investigated in strongly modified habitats such as arable lands (Clough et al., 2005; Schmidt et al., 2005, 2008; Öberg et al., 2007; Concepción et al., 2008; Drapela et al., 2008; Plass et al., 2010) and in mowed (grazed or unmowed) semi-natural grasslands (Hendrickx et al., 2007; Batáry et al., 2008, 2012; Miyashita et al., 2012; Zulka et al., 2014).

Surprisingly, the assessment of the importance of local and landscape-level factors on the assemblages of unmanaged grasslands is still missing. We would like to stress that there is an essential difference between the spider assemblages living in strongly modified, disturbed habitats, and those inhabiting unmanaged habitats. Assemblages in modified, disturbed habitats are regularly disrupted, therefore, recolonization from the surrounding landscape may be permanent, while assemblages in unmanaged habitats are more stable, and therefore, the role of the recolonization may be less important. Therefore, it is important to test the influence of the local and landscape-level factors in these unmanaged habitats.

In this study, our aim was to test the effects of local and landscape-level factors on the species richness and abundance. We also tested the effects of local and landscape-level factors on the abundance of the most frequent hunting spider species. We hypothesized that spider species richness and abundance increase with increasing cover of plants and litter, average height of grass and patch size but decrease with increasing bare ground. We supposed that in unmanaged grasslands, the spider assemblages are stable; therefore, the effects of landscape-level factors (isolation and landscape diversity) are less important in shaping spider assemblages than local factors. Moreover, we also hypothesized that most frequent hunting species respond heterogeneously to both local and landscape-level factors due to their variability in habitat affinity and ecological demands.

2. Methods

2.1. Study area

We selected nine unmanaged dry sandy grassland fragments in the Kiskunság region (size: 30,628 ha) of the Hungarian Great Plain which is located between the Danube and the Tisza rivers (Central Hungary) (Table 1). All selected sampling sites have the same vegetation type; these fragments are embedded in the same matrix. The typical grassland vegetation of the unmanaged dry sandy grassland fragments was Festucaea vaginatae (Szinétári et al., 2005). The relative abundance of protected plant species can exceed 26% in these grasslands (Török et al., 2000). Wetlands, forests (native and planted), and arable fields surrounded all the investigated grassland fragments. Thus, the habitat matrix was similar for all the studied grassland fragments. The Kiskunság region lies in the warm temperate zone with an annual mean temperature between 10.2 and 10.8 °C. The annual mean precipitation is 550–600 mm with two maxima in May and November and summer drought (Török et al., 2000). The most typical soil types are sand with more or less humus content and saline soils. The Kiskunság region is an important biodiversity hotspot in Hungary, because of its unique native plant and animal communities on sandy areas (Török et al., 2000; Szinétári et al., 2005; Batáry et al., 2007). The region is characterised by a mosaic of natural grasslands (sandy grasslands and salt meadows), wetlands (marshes, fen meadows, and mires), and forests (sandy oak woods, poplar-juniper steppe woodlands, and floodplain forests), as well as arable fields (maize and corn) and non-native tree plantations (black locust (Robinia pseudoacacia), ennobled poplar species (Populus spp.), and pine species (Pinus spp.)). After introducing arable farming in the 18th century, the area of the natural habitats decreased significantly by the end of the last century. Nowadays, only remnants of these habitats are to be found within the fragmented landscape of the Kiskunság region. The minimum distance between the studied grassland fragments was 1 km; the maximum distance was 44 km, while the average distance between the fragments was 22 km.

2.2. Sampling design

During the 9-year study period (2001–2009), we sampled spider species using pitfall traps. We placed 10 traps randomly in each investigated fragment. All traps were at least 50 m apart from the grassland edges to avoid edge effects (Horváth et al., 2002). Traps consisted of plastic cups with 100 mm diameter and contained about 150 ml 70% ethylene glycol as a killing-preserving

<table>
<thead>
<tr>
<th>Fragments/variables</th>
<th>Cover of plants (%)</th>
<th>Cover of bare ground (%)</th>
<th>Cover of litter (%)</th>
<th>Average height of grass (cm)</th>
<th>Patch size (ha)</th>
<th>Inverse isolation index (ha)</th>
<th>Landscape diversity</th>
</tr>
</thead>
<tbody>
<tr>
<td>1. Bugac-borókás</td>
<td>36.7</td>
<td>49.0</td>
<td>4.3</td>
<td>29.4</td>
<td>408.0</td>
<td>556.0</td>
<td>0.261</td>
</tr>
<tr>
<td>2. Bugac-legeló</td>
<td>72.8</td>
<td>3.5</td>
<td>16.0</td>
<td>34.6</td>
<td>1022.4</td>
<td>688.1</td>
<td>0.372</td>
</tr>
<tr>
<td>3. Fischerbócsa</td>
<td>46.9</td>
<td>36.9</td>
<td>7.1</td>
<td>27.7</td>
<td>827.9</td>
<td>80.6</td>
<td>0.170</td>
</tr>
<tr>
<td>4. Fülfóházá</td>
<td>54.9</td>
<td>32.3</td>
<td>4.1</td>
<td>32.0</td>
<td>300.8</td>
<td>526.9</td>
<td>0.524</td>
</tr>
<tr>
<td>5. Kunadacs</td>
<td>54.5</td>
<td>22.0</td>
<td>20.7</td>
<td>32.0</td>
<td>3.0</td>
<td>28.5</td>
<td>0.113</td>
</tr>
<tr>
<td>6. Kunbaracs east</td>
<td>63.0</td>
<td>13.5</td>
<td>29.1</td>
<td>14.6</td>
<td>30.0</td>
<td>63.2</td>
<td>0.263</td>
</tr>
<tr>
<td>7. Kunbaracs west</td>
<td>59.6</td>
<td>14.6</td>
<td>28.8</td>
<td>14.6</td>
<td>8.6</td>
<td>57.0</td>
<td>0.148</td>
</tr>
<tr>
<td>8. Orgovány</td>
<td>67.8</td>
<td>15.7</td>
<td>8.5</td>
<td>30.0</td>
<td>108.0</td>
<td>185.7</td>
<td>0.365</td>
</tr>
<tr>
<td>9. Soltszentimere</td>
<td>42.3</td>
<td>42.7</td>
<td>7.9</td>
<td>34.3</td>
<td>175.5</td>
<td>270.7</td>
<td>0.539</td>
</tr>
</tbody>
</table>
Liquid. We protected pitfall traps by fiberboard from litter, rain, and small vertebrates. We emptied the traps fortnightly from the end of March to the end of October. We identified all adult spiders to species level, while juvenile spiders only to genus level (Nentwig et al., 2015). We followed the World Spider Catalog’s nomenclature (World Spider Catalog, 2015).

2.3. Data analyses

Local and landscape-level factors affecting the distribution of spider assemblages were studied (Table 1). The local factors were the cover of plants, the cover of litter, the cover of bare ground, the average height of grass, and the patch size. We estimated these parameters in every year within a 1 × 1 m quadrant next to each trap. We used the average value of these factors from the nine years for the statistical analyses. The size of the studied unmanaged dry sandy grassland fragments was measured on digitized 1:10,000 maps and aerial photographs using Quantum GIS program. Two measures of landscape composition were analyzed: (i) inverse isolation index, and (ii) landscape diversity (Table 1). We measured these factors in the first (2001) and last (2009) years of the study. Isolation of a grassland fragment is often measured as the distance to the nearest fragment, notwithstanding isolation also depends on the size of the nearest fragment. We measured the isolation of the grassland fragments by the inverse isolation index, defined as the total size of unmanaged dry sandy grasslands within buffer around the studied grassland fragment. This value decreases as the isolation of the grassland increases (Magura et al., 2001). We also identified the different vegetation patches based on aerial photographs within buffer around the studied grassland fragment. To express the landscape diversity, we calculated the Shannon diversity index based on the area of six habitat types: grasslands, wet meadows, forests, arable fields, artificial areas, and water bodies. A series of buffers with increasing distances (100 m, 500 m, 1000 m, 2000 m, 3000 m) from the edge of the grassland fragments around all sites were used to determine appropriate spatial scale at which the surrounding landscape is relevant and influencing the spider species richness and abundance.

We investigated the relationships between the local and landscape-level factors and the species richness and abundance of spider assemblages by generalized linear models (GLMs, StatSoft Inc., 2010) using the multiple regression design. We first fitted the full model containing all (local and landscape-level) factors. We evaluated models based on Akaike’s information criterion (Akaike, 1973) and accepted the model with the lowest AIC as the final model. In the final model, the dependent variables (species richness and abundance) were regarded as following quasi-Poisson distribution (with log link function) to account for overdispersion in the data (Bolker et al., 2009). We pooled the catches of all traps for the whole year in every year and analyzed the nine years average value of all sites. The overall species richness, the total number of individuals of the species richness and abundance of habitat specialist, generalist and hunting spiders were regarded as dependent variables. We considered the following species as habitat specialists: (i) species which occur exclusively in open sandy habitats, (ii) species which occur in more open habitat types, but in lowlands can be found only in open sandy habitats. For all dependent variables, the goodness of fit of the model with the lowest AIC was the highest in case of the buffer with 1000 m distance from the edge of the grassland fragments; therefore, this spatial scale was used during the analyses. The number and the area of the unmanaged dry sandy grasslands within the buffer around the studied grassland fragments did not change between 2001 and 2009. Moreover, the landscape diversity did not differ significantly between the studied two years (Mann–Whitney U = 36.00, p = 0.6911, N = 9). Thus, we used the data from 2009 for the statistical analyses.

We examined the relationships between the local and landscape-level factors and the abundance of the nine most frequent species using the detrended canonical correspondence analysis by second order polynomials (DCCA) calculated by the CANOCO package (Leps and Smilauer, 2003). Biplot scaling in the ordination was focused on the inter-species distances.

3. Results

During the 9-year study, we collected altogether 6589 individuals of 145 species, including 46 habitat specialist species with 3008 individuals, 85 generalist species with 3314 individuals, 9 forest-associated species with 244 individuals and 5 species, which could be determined only at genus level with 23 individuals (Electronic supplementary material (ESM) Table 1). Regarding the main guild types, 6273 individuals belonged to 110 hunting species and 316 individuals represented 35 web-building species. Four of the nine most abundant species were habitat specialist (Allopecosa psammophila, Berlandia cinerea, Callilepis nocturna, and Gnaphosa montolica), four were generalist (Allopecosa cursor, Allopecosa sulzeri, Xysticus kochi, and Zelotes longipes) and one was forest-associated species (Pardosa alatricis).

The generalized linear models showed that the local factors (cover of plants, cover of litter cover of bare ground, and patch size) had more influence on spider assemblages, than the landscape-level factors (Figs. 1–3, Tables 2 and 3). The relationship was significantly positive between the total number of species, the number of habitat specialist species, the number of habitat specialist individuals, the number of generalist individuals, and the cover of plants (Fig. 1a–d, Table 2). A significant positive relationship was found between the number of individual generalist individuals, the number of specialist individuals, the number of hunting species, and the cover of litter (Fig. 2a–c, Tables 2 and 3). The total number of species, the total number of individuals, the number of generalist species, the number of generalist individuals, and the number of hunting individuals decreased significantly with the increasing of the cover of bare ground (Fig. 3a–e, Tables 2 and 3). The number of generalist and specialist individuals increased significantly with the increasing of the patch size (Table 2). Total number of individuals and number of hunting individuals showed a significant negative, while the number of specialist individuals showed a significant positive relationship with the inverse isolation index (Tables 2 and 3). One studied local factor (average height of grass) and one landscape-level factor (landscape diversity) had no significant influence on any of the dependent variables (Tables 2 and 3).

The DCCA biplot showed that there was a negative relationship between the cover of bare ground and the cover of plants (Spearman’s rank correlation coefficient r = −0.9620) and the cover of litter (Spearman’s rank correlation coefficient r = −0.6966) (Fig. 4). The response of individual hunting species to both local and landscape-level factors was heterogeneous. The habitat specialist A. psammophila was positively associated with bare ground cover and negatively with plant and litter cover, while the also specialist B. cinerea showed an opposite reaction to these factors (Fig. 4). Neither local nor landscape-level factors influenced the distribution of the generalist species (A. cursor, A. sulzeri, X. kochi, and Z. longipes). Similarly, the distribution of a specialist (C. nocturna) and a forest-associated species (P. alatricis) was not influence neither the local nor the landscape-level factors (Fig. 4). The specialist C. montolica was the only species that responded merely to one landscape-level factor. The abundance of this species increased with the decreasing of isolation (Fig. 4).
Fig. 1. Relationship between the cover of plants and the total number of species (a), the number of specialist species (b), the number of specialist individuals (c), and the number of generalist individuals (d). Dashed lines represent the confidence bands (95%).

Fig. 2. Relationship between the cover of litter and the number of generalist individuals (a), the number of specialist individuals (b), and the number of hunting species (c). Dashed lines represent the confidence bands (95%).
Fig. 3. Relationship between the cover of bare ground and the total number of species (a), the total number of individuals (b), the number of generalist species (c), the number of generalist individuals (d), and the number of hunting individuals (e). Dashed lines represent the confidence bands (95%).

Table 2

| Relationship between the number of spider individuals, species and the studied factors by generalized linear models (GLMs) using the multiple regression design. |
|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|
| Total number of individuals | Total number of species | Number of specialist individuals | Number of specialist species | Number of generalist individuals | Number of generalist species |
| Cover of plants | Not entered | *** | Not entered | *** | Not entered | *** |
| Cover of litter | Not entered | *** | Not entered | *** | Not entered | *** |
| Cover of bare ground | Not entered | *** | Not entered | *** | Not entered | *** |
| Average height of grass | Not entered | Not entered | Not entered | Not entered | Not entered | Not entered |
| Patch size | Not entered | Not entered | *** | Not entered | *** | Not entered |
| Inverse isolation index | Not entered | Not entered | *** | Not entered | *** | Not entered |
| Landscape diversity | Not entered | Not entered | Not entered | Not entered | Not entered | Not entered |

Not significant (ns), significant negative (−) and significant positive (+) relationships are marked. Not entered: the factors were not entered into the final model based on Akaike's information criterion (AIC). *p < 0.05, **p < 0.01, ***p < 0.001
Table 3
Relationship between the number of hunting spider individuals, species and the studied factors by generalized linear models (GLMs) using the multiple regression design.

<table>
<thead>
<tr>
<th>Factor</th>
<th>Number of hunting individuals</th>
<th>Number of hunting species</th>
</tr>
</thead>
<tbody>
<tr>
<td>Cover of plants</td>
<td>Not entered</td>
<td>Not entered</td>
</tr>
<tr>
<td>Cover of litter</td>
<td>***</td>
<td>Not entered</td>
</tr>
<tr>
<td>Cover of bare ground</td>
<td>Not entered</td>
<td>Not entered</td>
</tr>
<tr>
<td>Average height of grass</td>
<td>Not entered</td>
<td>Not entered</td>
</tr>
<tr>
<td>Patch size</td>
<td>Not entered</td>
<td>Not entered</td>
</tr>
<tr>
<td>Inverse isolation index</td>
<td>–*</td>
<td>Not entered</td>
</tr>
<tr>
<td>Landscape diversity</td>
<td>Not entered</td>
<td>Not entered</td>
</tr>
</tbody>
</table>

Not significant (ns), significant negative (–) and significant positive (+) relationships are marked. Not entered: the factors were not entered into the final model based on Akaike’s information criterion (AIC). *p < 0.05, **p < 0.01, ***p < 0.001.

4. Discussion

In unmanaged dry sandy grassland fragments, we investigated the effects of local and landscape-level factors on spider assemblages. We found that the local factors had stronger influence on species richness and density of spiders than landscape-level factors. Only one landscape-level factor (isolation) had effects on the total abundance and on the abundance of specialist and hunting species.

The local factors (four out of five: plant cover, litter cover, bare ground cover, and patch size) significantly affected overall, generalist, specialist, and hunting spider assemblages. The species richness and abundance were positively influenced by the plant cover, litter cover and patch size, but negatively by the bare ground cover. There are several retreats, hiding places, and numerous preys in a dense vegetation and under litter; these results in a higher species richness and abundance of spiders in the patches with higher plant and litter cover. Furthermore, the increasing plant and litter cover moderate temperature and humidity extremes contributing to the enhancement of spider diversity (Rypstra et al., 1999). Similarly to our findings Batáry et al. (2008) found a positive effect of plant and litter cover on species richness and abundance of spiders. In contrast, Warui et al. (2005) found no significant relationship between the diversity of ground-dwelling spiders and the cover of plants, studying the effects of grazing by cattle, mose-herbivores like buffaloes and other smaller ungulates, and mega-herbivores like giraffes and elephants. During the analysis, the effect of herbivore species was not considered; the contrasting effect of smaller ungulates, mose- and mega-herbivores may annihilate the influence of plants’ cover on spider diversity resulting in a spurious no influence. Dennis et al. (2001) studied the epigeal arachnids of *Nardus stricta* dominated grassland in Scotland and they emphasized that the vegetation structure (mean vegetation height) was the most important local factor which significantly influenced the species composition and abundance. Our species-level analysis showed that the specialist nocturnal hunter *B. cincta*, which prefers the open and dry sandy grasslands (Buchar and Ruzicka, 2002) reacted positively to plant and litter cover and negatively to bare ground cover. This spider needs some vegetation including lichens, and litter as hiding place, because this species takes shelter among vegetation and under litter not in the soil in the daytime (Buchar and Ruzicka, 2002).

We found that the bare ground cover decreased the diversity of the spider assemblages. Most of spiders have conspicuous coloration and markings, and they can not find suitable shelter away from predators if the soil surface becomes open. The abundance and species richness of these species decrease with the increase of bare ground cover (Wu et al., 2009). The intense management in grasslands increases the cover of bare ground, which causes the decrease of the diversity of spider assemblages and the increase of the dominance of a few r-selected species; this initial colonizing species are very mobile and less demanding toward habitat features (Bell et al., 2001). In our study, the abundance of sand specialist *A. psammophila* positively correlated with bare ground cover and negatively with plant and litter cover. Our findings suggest that this species occurs in high number only in those fragments where the ratio of the bare sand surfaces is considerably high.

Patch size influenced both the number of generalist and specialist abundance. The abundance of these spiders increased with the patch size. Gallé (2008) also reported that the abundance of generalist and specialist spiders increased with patch size. It is likely that both generalist and specialist species need a minimum habitat size for a viable population. The abundance of these species decreased with the decrease of patch size suggesting that they were not able to settle down permanently in the smaller fragments. It is likely that, these spiders can survive and reproduce with major chance only in the larger grassland fragments.

We found that isolation was the only landscape-level factor which influenced the abundance of spiders. The overall spider abundance and the abundance of hunting species increased, while the abundance of specialist species and the abundance of specialist *G. mongolica* decreased as isolation of the grassland fragments increased. The neighboring habitats around the more isolated fragments may function as the source habitats in case of generalist, forest-associated and hunting species; therefore, these spiders can colonize the more isolated unmanaged dry sandy grassland fragments from the agricultural and forested lands. However, habitat specialist spiders could not settle down permanently in the adjacent arable lands and non-native tree plantations, because these habitats were unfavorable for them. These habitats functioned as barriers for specialist species, thus, their abundance...
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Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, at http://dx.doi.org/10.1016/j.agee.2015.04.033.

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